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CHAPTER 16. MEASUREMENT OF ATMOSPHERIC COMPOSITION**16.1 GENERAL**

The main purpose of this chapter is to introduce readers (particularly those who are new to these types of measurements) to methods and specific techniques used for measuring various components of atmospheric composition and a number of related physical parameters. This is often accompanied by measurements of basic meteorological variables, as introduced in the preceding chapters. Within WMO, the Global Atmosphere Watch (GAW) Programme was established in response to the growing concerns related to human impacts on atmospheric composition and the connection of atmospheric composition to weather and climate. GAW's mission is focused on the systematic global observations of the chemical composition and related physical characteristics of the atmosphere, integrated analysis of these observations and development of predictive capacity to forecast future atmospheric composition changes (WMO, 2017a).

The observations and analyses of the atmospheric chemical composition are needed to advance the scientific understanding of the effects of the increasing influence of human activity on the global atmosphere, as illustrated by pressing societal problems such as: changes in the weather and climate related to human influence on atmospheric composition, particularly, on greenhouse gases, ozone and aerosols; impacts of air pollution on human and ecosystem health and issues involving long-range transport and deposition of air pollution; changes in UV radiation as consequences of changes in atmospheric ozone amounts and climate, and the subsequent impact of these changes on human health and ecosystems.

For further practical details on measurement activities, see the GAW reports and other references listed at the end of the chapter.

The GAW observations focus on six classes of variables:

- (a) Ozone: column (total) ozone and ozone vertical profiles with a focus on the stratosphere and upper troposphere;
- (b) Greenhouse gases: carbon dioxide CO₂ (including $\Delta^{14}\text{C}$, $\delta^{13}\text{C}$ and $\delta^{18}\text{O}$ in CO₂, and oxygen/nitrogen (O₂/N₂) ratios), methane CH₄ (including $\delta^{13}\text{C}$ and δD in CH₄), nitrous oxide (N₂O) and halogenated compounds (SF₆);
- (c) Reactive gases: surface and tropospheric ozone (O₃), carbon monoxide (CO), volatile organic compounds (VOCs), nitrogen oxides (NO_x), sulphur dioxide (SO₂) and molecular hydrogen (H₂), ammonia (NH₃)¹;
- (d) Atmospheric total² deposition (focused largely on major ions in the wet deposition group);

¹ Ammonia was identified as one of the key substances required to address the nitrogen cycle but the recommendations concerning measurement guidelines are not yet available.

² Measurement techniques for dry deposition have not been recommended yet.

- (e) Ultraviolet radiation;
- (f) Aerosols (including near-surface in-situ physical, optical and chemical properties, total integrated column properties and profiling).

A number of ancillary parameters are recommended for measurement at GAW stations:

- (a) Solar radiation;
- (b) Major meteorological parameters;
- (c) Natural radioactivity including krypton-85, radon and some other radionuclides.

Atmospheric water vapour was tentatively included in GAW in 2015 at the decision of the Environmental Pollution and Atmospheric Chemistry Scientific Steering Committee (WMO, 2015a) but the infrastructure has yet to be defined.

Due to the low mixing ratios of atmospheric trace constituents, the instruments and methods used for the quantitative and qualitative determination of atmospheric constituents are complex and sometimes difficult to operate. Small errors, for example in spectral signatures or cross-sensitivities to other compounds, can easily confound the accuracy of atmospheric composition measurements. Therefore, besides correct operation, regular calibration of the equipment, participation in intercomparison exercises, station audits and personnel training are essential for accurate and reliable measurements. Obtaining reliable and high-quality results for most of the measurements described here is not feasible without the close involvement of specialist staff at a professional level. The main principles of the quality assurance of atmospheric composition observations within GAW are described in 16.1.4.

The GAW Programme Implementation Plan builds around the concept of "science for services" and addresses multiple applications that utilize atmospheric chemical composition observations. The need to improve knowledge of air quality variations, for example, on an urban scale, along with current technical development, has brought new measurement techniques into the community, namely low-cost environmental sensors. Low-cost air pollution sensors have the potential to deliver new information, but they are also prone to a number of shortcomings. Long-term suitability of such sensors has yet to be proven and an ongoing broader assessment reviews different sensor technologies and application-specific requirements for data quality and calibration. As this is a fast-changing field, continuous re-evaluation, including new developments and changes in performance, may be required. An expert group under the umbrella of the Commission for Atmospheric Sciences is currently developing a comprehensive statement on the use of low-cost sensor technology.

16.1.1 Definitions/descriptions

Depending on the measurement principle and instrument platform, two types of measurements are routinely performed and reported, namely:

- (a) point measurements;
- (b) integrated measurements.

"Point measurements" refers to the results of (continuous or discrete) measurements of a particular component's quantity in a specific place in space (either in an atmospheric layer of a few tens of metres above the surface at a particular location on the Earth's surface or anywhere in the troposphere, the stratosphere or any other atmospheric layer). A series of point measurements at several altitudes above a given location constitute a vertical profile measurement (for example, measurements from an aircraft or balloons/sondes, rockets etc.). The point measurements can also be performed along specific horizontal routes using mobile platforms (for example, ship, train, car). Results of point measurements are commonly given in units of partial pressure, concentration, mixing ratio or mole fraction, corrected for standard temperature and pressure conditions of 273.15 K and 101.325 kPa, respectively, when measurements are performed above sea-

level. The use of units that are not part of the International System of Units (SI) is strongly discouraged.

“Integrated measurements” refers to the integrated or average amount of a particular substance contained in the atmosphere along the observational path. This can be a vertical total column extending from the Earth’s surface to the upper edge of the atmosphere. Commonly used units of total ozone are (i) equivalent column thickness of a layer of pure ozone at standard temperature and pressure conditions of 273.15 K and 101.325 kPa, respectively, and (ii) vertical column density (total number of molecules per unit area in an atmospheric column). For the other atmospheric constituents, vertical column density or column-averaged abundances are used. It is also common to report the partial column content of a substance, for example the tropospheric column content of NO_x. Here, the vertical column that is integrated extends from the Earth’s surface to the tropopause. The differential optical absorption spectroscopy (DOAS) instruments also allow for measurements of average amount of substance along horizontal pathways.

Observations of atmospheric composition include gaseous composition, aerosol and total atmospheric deposition. The characteristics of precipitation chemical composition (wet deposition) are given in 16.5. The variables describing aerosols (physical and chemical properties) are listed in 16.6.

16.1.2 Units and scales

The following units are used to express the results of atmospheric trace compound observations:

Number of molecules per unit area: represents the column abundance of atmospheric trace compounds. Still widely used is the Dobson unit (DU), which corresponds to the number of molecules of ozone required to create a layer of pure ozone 10⁻⁵ m thick at standard temperature and pressure (STP). Expressed another way, 1 DU represents a column of air containing about $2.6868 \cdot 10^{16}$ ozone molecules for every square centimetre of area at the base of the column.

Mass concentration: represents the mass of the specific constituent in unit volume of the atmospheric air (for example, microgram per m³).

Milliatmosphere centimetre (m-atm-cm): A measure of total ozone equal to a thickness of 10⁻³ cm of pure ozone at STP (1 m-atm-cm is equivalent to 1 DU).

Mole fractions of substances in dry air (dry air includes all gaseous species except water vapour (H₂O)):

$$\begin{aligned}\mu\text{mol mol}^{-1} &= 10^{-6} \text{ mole of trace substance per mole of dry air} \\ \text{nmol mol}^{-1} &= 10^{-9} \text{ mole of trace substance per mole of dry air} \\ \text{pmol mol}^{-1} &= 10^{-12} \text{ mole of trace substance per mole of dry air}\end{aligned}$$

Dry mole fraction requires either drying of air samples prior to measurement or correction of the measurement for water vapour abundance. When drying is impossible or the correction would add substantial uncertainty to the measurement, wet mole fractions can be reported instead. This must be clearly indicated in the metadata of the observational record.

The appropriate unit for expressing amount of substance is dry-air mole fraction, reported as ppm (parts per million, i.e. $\mu\text{mol mol}^{-1}$), ppb (parts per billion, i.e. nmol mol^{-1}) or ppt (parts per trillion, i.e. pmol mol^{-1}). A “v” has often been appended to these units to indicate mixing ratio by volume. When reporting mole fractions as volume mixing ratios, one assumes the atmosphere to be an ideal gas. Deviations from the ideal under GAW conditions can be large (such as for CO₂), so the use of mole fraction is strongly preferred because it does not require an implicit assumption of ideality of the gases and, more importantly, because it is also applicable to condensed-phase species. In general, the use of SI units is highly recommended.

Isotope or molecular ratios:

Atmospheric molecules can be present in different isotopic configurations.³ Isotope ratio data are expressed as deviations from an agreed upon reference standard using the delta notation:

$$\delta = \left(R_{\text{sample}} / R_{\text{reference}} - 1 \right), \text{ with } R = [\text{heavy isotope}] / [\text{light isotope}] \quad (16.1)$$

δ -Values are expressed in multiples of 1 000 (‰ or per mil).

The international reference scale (i.e. the primary scale) for $\delta^{13}\text{C}$ is Vienna Pee Dee Belemnite (VPDB). NBS 19 and LSVEC (Coplen et al., 2006) are the primary international reference materials defining the VPDB scale. For $\delta^{18}\text{O}$, multiple scales are in use (VPDB, Vienna Standard Mean Ocean Water (VSMOW), air- O_2 (de Laeter et al., 2003)).

The delta notation is also used to express relative abundance variations of O_2/N_2 (and argon/nitrogen (Ar/N_2)) ratios in air:

$$\delta(\text{O}_2/\text{N}_2) = \left(R_{\text{sample}} / R_{\text{standard}} - 1 \right), \text{ with } R = \text{O}_2/\text{N}_2 \quad (16.2)$$

The respective international air standard is not yet established. The Scripps Institution of Oceanography (SIO) local O_2/N_2 scale, based on a set of cylinders filled at the Scripps Pier, is the most widely used scale.

$\delta(\text{O}_2/\text{N}_2)$ values are expressed in multiples of 10^6 or per meg.

Precipitation chemistry (wet deposition) observations include measurements of several parameters that are described in more detail in 16.5. The following units are used:

- pH measurements are expressed in units of acidity defined as: $\text{pH} = -\log_{10} [\text{H}^+]$, where $[\text{H}^+]$ is expressed in mole L^{-1} ;
- Conductivity is expressed in $\mu\text{S cm}^{-1}$ (microsiemens per centimetre), a unit commonly used for measuring electric conductivity;
- Acidity/alkalinity is expressed in $\mu\text{mole L}^{-1}$ (micromole per litre);
- Major ions content is expressed in mg L^{-1} (milligram per litre).

Aerosol observations of volumetric quantities, i.e. the amount of substance in a volume of air, are reported for STP. These may refer to a particle number concentration (cm^{-3}), an area concentration ($\text{m}^2 \text{m}^{-3}$, or m^{-1}) or a mass concentration ($\mu\text{g m}^{-3}$). Aerosol optical depth is a dimensionless quantity. Absorption and scattering coefficients are expressed in m^{-1} .

16.1.3 Measurement principles and techniques

The existing techniques for atmospheric chemical composition measurements can be separated into three main groups: passive sampling, active sampling and remote-sensing techniques, which can themselves be both active (for example, lidar systems with their own light source) and passive (spectrometers using, for example, sunlight). Essentially, active sampling techniques draw the air sample through the detector or sampling device by a pump, whereas passive sampling techniques use the diffusion of air to the sampling device. In remote-sensing techniques, the ana-

³ CO_2 , for example, mostly consists of $^{12}\text{C}^{16}\text{O}^{16}\text{O}$, while the smaller abundance higher-mass isotopologues from mass 45 up to mass 49 ($^{13}\text{C}^{16}\text{O}^{16}\text{O}$, $^{14}\text{C}^{16}\text{O}^{16}\text{O}$, or $^{12}\text{C}^{18}\text{O}^{16}\text{O}$, the corresponding ^{17}O siblings and the mixed-isotope species) are also found in the atmosphere.

lysed air volume and the detector are at different locations. Total or partial column measurements are possible only with remote-sensing techniques.

In the case of active sampling, measurements can either be made continuously (or at least quasi-continuously with short integration times)⁴ or samples can be collected or specially prepared (in glass or stainless steel cylinders, on sorbent substrates or filters) and analysed offline in specialized laboratories. The collection of discrete samples entails the storage of these samples. During this time, flask properties may influence the composition of the sample due to chemical or surface effects or permeation through sealing polymers. This demands careful tests of the sampling containers.

The analytical techniques most commonly used (and recommended in the GAW Programme) for detecting and quantifying atmospheric trace constituents can be summarized as follows:

- (a) *Spectroscopic methods* refer to the measurement of changes in radiation intensity due to absorption, emission, photoconductivity or Raman scattering by a molecule or aerosol particle as a function of wavelength. Spectral measurement devices are referred to as spectrometers, spectrophotometers, spectrographs or spectral analysers. Spectral measurements can be performed in different parts of a spectrum depending on the component to be measured, or at several individual wavelengths. As absorption lines are different for molecules with different isotopic composition, and line shapes depend on the bulk composition of the gas, care should be taken to ensure that reference gases have similar properties to the analysed atmospheric air.
- (b) *Gas chromatography* (GC) is a physical method of separation that distributes components to be separated between two phases, one stationary (stationary phase), the other (the mobile phase) moving in a definite direction. There are numerous chromatographic techniques and corresponding instruments. To be suitable for GC analysis, a compound must have sufficient volatility and thermal stability. Gas chromatography involves a sample being vapourised and injected onto the head of the chromatographic column. The sample is transported through the column by the flow of an inert, gaseous mobile phase. The column itself contains a liquid stationary phase which is adsorbed onto the surface of an inert solid. A chromatography detector is a device used to visualize components of the mixture being eluted off the chromatography column. There are two general types of detectors: destructive and non-destructive. The destructive detectors, such as a flame ionization detector (FID), perform continuous transformation of the column effluent (burning, evaporation or mixing with reagents) with subsequent measurement of some physical property of the resulting material (plasma, aerosol or reaction mixture). The non-destructive detectors, such as an electron capture detector (ECD), are directly measuring some property of the column effluent (for example UV absorption) and thus allow for further analyte recovery.
- (c) *Mass spectrometry* (MS) is an analytical technique that produces spectra of the masses of the molecules comprising a sample of material. The spectra are used to determine the elemental composition of a sample, the masses of particles and of molecules, and to elucidate the chemical structures of molecules. Mass spectrometry works by ionizing chemical compounds to generate charged molecules or molecule fragments and measuring their mass-to-charge ratios. In a number of instruments mass spectrometry can be used as a detector method for gas chromatography.

Some other analytical techniques can be found in use at GAW stations where the amount of the analyte is defined through its chemical reaction with the working body (for example, electrochemical method or methods based on chemiluminescence).

⁴ This is, for example, common practice in gas chromatography measurements.

Detection methods of gases and aerosols vary and are based on different physical phenomena. Details of the detection methods applicable to different gases and aerosol properties are summarized in the sections below.

The measurement techniques for the main compounds observed under the GAW Programme are briefly described in this chapter, while comprehensive measurement guidelines can be found in the specialized GAW reports, cited in individual sections. In the cases where GAW measurement guidelines or standard operating procedures are not available, links are provided to the information necessary to carry out the respective measurements.

Satellite remote-sensing of the atmospheric species mentioned below is treated separately in Volume IV, Chapter 5.

16.1.4 Quality assurance

The objectives of the GAW quality assurance (QA) system are to ensure that data reported by station operators are consistent, of known and adequate quality, supported by comprehensive metadata, and regionally or globally representative with respect to spatial and temporal distribution.

The principles of the GAW QA system (WMO, 2017a) apply to each measured variable and include:

- (a) Full support of the GCOS Climate Monitoring Principles.
- (b) Network-wide use of only one reference standard or scale (primary standard). As a consequence, only one institution is responsible for this standard.
- (c) Full traceability to the primary standard of all measurements made by Global, Regional and Local GAW stations and network standards of contributing networks where such standards are established.
- (d) The definition of data quality objectives (DQOs).
- (e) Establishment of guidelines on how to meet these quality targets, i.e. harmonized measurement techniques published as Measurement Guidelines (MGs) and Standard Operating Procedures (SOPs) and implemented at the stations.
- (f) Use of detailed log books for each parameter containing comprehensive meta information related to the measurements, maintenance, and 'internal' calibrations.
- (g) Regular independent assessments (system and performance audits).
- (h) Timely submission of data and associated metadata to the responsible World Data Centre or a contributing network data centre as a means of permitting independent review of data by a wider community.
- (i) Regular statistical and scientific analysis of data in the GAW data archives to ensure correctness, long-term consistency, and comparability of the archived measurement data.

Moreover, the Global Climate Observing System (GCOS) monitoring principles (WMO, 2016a) also apply to the GAW observations. Among those, the most relevant principles for atmospheric composition measurements are:

- (a) The impact of new systems or changes to existing systems should be assessed prior to implementation;
- (b) A suitable period of overlap for new and old observing systems should be required;
- (c) Operation of historically-uninterrupted stations and observing systems should be maintained.

The GAW QA system further recommends the adoption and use of internationally accepted methods and vocabulary to describe uncertainty in measurements.

Five types of central facilities (see Annex) dedicated to the six groups of measurement variables (see 16.1) are operated by WMO Members and form the basis of the quality assurance and data archiving system. These include:

- (a) Central Calibration Laboratories (CCLs), which host primary standards and scales;
- (b) World/Regional Calibration Centres (WCCs/RCCs), which coordinate intercomparison campaigns, help with instrument calibration and perform station/laboratory audits;
- (c) Quality Assurance/Science Activity Centres (QA/SACs), which provide technical and scientific support and coordinate cooperation between the central facilities and GAW stations;
- (d) World Data Centres (WDCs), which mainly ensure dissemination and easy access of GAW data and secure the data through appropriate data archiving.

The nature of the data flow and the share of responsibilities for QA/QC implementation between the Central Facilities and the observation sites is specific to each variable. The work of the central facilities on the quality assurance of the GAW observations is supported by respective scientific advisory groups, whose tasks include assisting in the development of measurement procedures and guidelines, data quality objectives and, when applicable, standard operating procedures, reviewing new measurement techniques and making recommendations about their applicability for the GAW observations. See the GAW Implementation Plan 2016-2023 (WMO, 2017a) for more details and the terms of reference for the central facilities and the scientific advisory groups. An important issue is the recent establishment in Europe of Atmospheric Research Infrastructures such as Integrated Carbon Observing System (ICOS, <https://www.icos-ri.eu>) and Aerosol, Cloud and Trace Gases Research Infrastructure (ACTRIS, <https://www.actris.eu>) that are implementing Central Facilities for quality control and quality analysis in support of GAW QA/QC implementation strategies.

Given the variety in maturity of measurement and QC/QA systems within GAW it is particularly important that all measurements should be reported with an associated measurement uncertainty (ISO, 2008)

16.2 (STRATOSPHERIC) OZONE MEASUREMENTS

16.2.1 Ozone total column

Measurements of total ozone are possible using passive remote-sensing techniques only. The most precise information on total ozone and its changes at individual sites can be obtained by measurements from the ground, for example by solar spectroscopy in the 300–340 nm wavelength region. Within the GAW Programme, Dobson spectrophotometers (designed for manual operation) and Brewer spectrophotometers (designed for automatic operation) are used as the instruments of choice for routine total ozone observations, thus providing two independent networks.

Details of the total ozone measurements with the Dobson spectrometer and their quality assurance are provided in WMO (2008a). Total ozone observations are made with this instrument by measuring the relative intensities of selected pairs of ultraviolet wavelengths, called the A, B*, C, C', and D wavelength pairs, emanating from the sun, moon or zenith sky. The A wavelength pair,

for example, consists of the 3 055 Å (Ångström units, 1 Å = 0.1 nm) wavelength that is highly absorbed by ozone, and the more intense 3 254 Å wavelength that is relatively unaffected by ozone. Outside the Earth's atmosphere, the relative intensity of these two wavelengths remains essentially fixed. In passing through the atmosphere to the instrument, however, both wavelengths lose intensity because of scattering of the light by air molecules and dust particles; additionally, the 3 055 Å wavelength is strongly attenuated while passing through the ozone layer whereas the attenuation of the 3 254 Å wavelength is relatively weak. Therefore, the relative in-

tensity of the A wavelength pair as seen by the instrument varies with the amount of ozone present in the atmosphere since, as the ozone amount increases, the observed intensity of the 3 055 Å wavelength decreases, whereas the intensity of the 3 254 Å wavelength remains practically unaltered. Thus, by measuring the relative intensities of suitably selected wavelength pairs with the Dobson instrument, it is possible to determine how much ozone is present in a vertical column of air extending from ground level to the top of the atmosphere in the neighbourhood of the instrument. The result is expressed in terms of thickness of an equivalent layer of pure ozone at STP.

The measurement principle of the Brewer spectrophotometer is similar to that of the Dobson instrument. The operating procedures are provided by the manufacturing company at <http://www.kippzonen.com/?productgroup/26142/Brewer+Spectrophotometer.aspx>. The recommendations for the GAW network are available from the World Ozone and Ultraviolet Radiation Data Centre (WOUDC) at http://woudc.org/archive/Documentation/SOP_Documents/brewerspectrophotometer_sop-june2008.pdf.

Results of comparisons of Brewer and Dobson instruments, as well as recommendations on the operation of the Brewer instruments, are provided in the reports of the biennial WMO consultations on Brewer ozone and UV spectrophotometer operation, calibration and data processing (for example, see WMO, 2008b; WMO, 2015b; WMO, 2015c; WMO, 2016b).

The world (primary) standard instruments of Brewer and Dobson networks are calibrated by the Langley plot method performed at the Mauna Loa Observatory in Hawaii (every 2–4 years); standards used by the Regional Calibration Centres to propagate traceability are calibrated against the primary standard every 2–3 years; and the station instruments are calibrated by side-by-side comparison with the standard instruments every six years for the Dobson and two years for the Brewer spectrophotometers. In addition, three successful Langley plot campaigns at the Izaña Atmospheric Observatory, on the island of Tenerife, with primary and regional standard Dobson instruments have proved the suitability of that location and facility for this absolute calibration method.

Complementary measurements of total ozone are provided by the differential optical absorption spectroscopy (DOAS) type UV/visible spectrometers that also allow detection of various minor trace gases (such as NO₂ and BrO). The French instrument is called *Système d'Analyse par Observations Zénithales (SAOZ)*, but it is based on the same principle as DOAS. These instruments are part of the measurement suites within the Network for the Detection of Atmospheric Composition Change (NDACC, <http://www.ndsc.ncep.noaa.gov/instr/>). Compared to the more established Brewer/Dobson network, the measurement and analysis procedures for DOAS type instruments are less standardized, but regular comparison campaigns have been carried out. Other instruments providing total ozone measurements from the ground (Russian filter instruments or those of the DOAS/SAOZ type) are not operated under the same data QA/QC programme as Dobson and Brewer instruments. The Russian and other filter instruments are not independently calibrated, but are tied to either Dobson or Brewer instruments. Data quality of all individual total ozone series deposited at WOUDC needs to be documented for the users.

Determining ozone amounts by the DOAS technique relies on accurate knowledge of the absorption cross-section, determined in the laboratory. However, the absorption cross-sections are both temperature and pressure dependent and so not a constant in the atmosphere. Care should be taken to select the appropriate version of the cross-sections and the most recent data should be used (WMO, 2015d).

16.2.2 Ozone profile measurements

Measurements of the vertical ozone distribution are possible by both active and remote-sensing methods.

16.2.2.1 Umkehr method

Dobson and Brewer spectrometers can be used for the measurement of vertical ozone distribution utilizing the Umkehr method (WMO, 2008a; Jaroslowski, 2013). The reduction of the Umkehr measurement to an ozone profile requires a complex algorithm that includes knowledge of the radiative properties of the real atmosphere. As this knowledge changes, the algorithm will change. A standard Umkehr observation consists of a series of C-pair wavelength measurements made on a clear zenith sky during the morning or afternoon. The measurements are commenced a few minutes before sunrise and continued until the sun is at an elevation not lower than approximately 20 degrees, or commenced in the afternoon when the sun is at an elevation not lower than approximately 20 degrees and continued until shortly after sunset. The zenith sky must be free from clouds for a period of 30 min to 1 h near sunrise or sunset. This is especially true at low latitude stations where the sun rises or sets rapidly. At other times, it is desirable that the zenith sky be cloudless, but it is permissible that clouds cross it periodically when measurements are not being made. Umkehr observations cannot be made at a polar station or at high latitude stations during summertime when the sun does not sink below the horizon.

To be able to compute the vertical distribution of ozone, it is necessary to know the total amount of ozone present at the time of observation. Several total ozone measurements must, therefore, be made during the morning or afternoon, particularly if the ozone amount is changing fairly rapidly.

The resulting ozone profile derived from reduction of these measurements is quite dependent on the algorithm used. The Umkehr data analysis method was originally developed by Götz et al. (1934). Later the method was refined by Ramanathan and Dave (1957), Mateer and Dütsch (1964), and Mateer and DeLuisi (1992). The Umkehr algorithm is described by Petropavlovskikh et al. (2005), and updated information is available from <http://www.esrl.noaa.gov/gmd/ozwv/umkehr/>.

16.2.2.2 Ozonesonde measurements

Ozone measurement from balloons (ozonesondes) is an active method for measuring ozone's vertical distribution in the atmosphere. Other active methods for ozone mole fraction measurements (used on aircraft platforms) are described in the section on reactive gases (see 16.4.1).

Ozonesondes are small, lightweight and compact balloon-borne instruments, developed for measuring the vertical distribution of atmospheric ozone up to an altitude of about 30–35 km. The sensing device is interfaced to a standard meteorological radiosonde for data transmission to the ground station. Three major types of ozonesondes – the Brewer-Mast sonde, the electrochemical concentration cell (ECC) and the carbon iodine cell – are currently in use. Each sonde type has its own specific design.

The flight package typically weighs approximately 1 kg in total and can be flown on small weather balloons. Normally data are taken during ascent, at a vertical ascent rate of approximately 5 m/s, to a balloon burst altitude of 30–35 km. The inherent response time of the ozonesonde is 20–30 s such that the effective height resolution of the measured vertical ozone profile is typically 100–150 m.

The principles of ozonesonde operations and an overview of the different aspects of quality assurance and quality control for ozonesonde measurements in GAW are given in detail in WMO (2014).

16.2.2.3 Other measurement techniques

Ozone profile measurements can also be obtained by other instruments operated under the umbrella of NDACC. Lidar and microwave measurements are part of the NDACC suite of measurements and are valuable for assessing ozone trends in the upper stratosphere and for validating satellite measurements in the upper atmosphere. The disadvantage of microwave ozone measurements is the rather poor vertical resolution, but they have the potential to measure up to the mesopause region. The combination of sonde, Umkehr, lidar and microwave data from the ground

is important for assessing the quality of the ozone profile measurements from space (van der A et al., 2010).

16.2.3 Aircraft and satellite observations

Ozone in the atmosphere is also measured by instruments located on board aircraft and space-borne satellites. The airborne observations are usually made by in-situ photometers sampling the air in the troposphere and lower stratosphere during a flight. The measurements are used mostly in research campaigns on atmospheric chemistry, but there have also been long-term projects using commercial aircraft, such as MOZAIC (measurement of ozone, water vapour, carbon monoxide and nitrogen oxides aboard airbus in-service aircraft), CARIBIC (Civil Aircraft for the Regular Investigation of the Atmosphere Based on an Instrument Container, <http://www.caribic-atmospheric.com/>), and IAGOS (In-service Aircraft for a Global Observing System, <https://www.iagos.org/>).

Large-scale monitoring of atmospheric ozone is performed by remote-sensing instruments from satellites. These programmes can be divided according to lifetime: the long-term operational monitoring systems that generate large (global) datasets used for trend analyses and for operational mapping of ozone, and the temporary experimental missions.

Satellite observations can be grouped according to the radiation-detection technology used for the instruments and the retrieval schemes applied for the derivation of ozone column density or concentration from the measured radiances. While nadir-viewing instruments are primarily used for column observations and coarse vertical profiling, limb sounding instruments are able to measure vertical profiles of ozone at high vertical resolution by solar, lunar or stellar occultation or by observing limb scatter and emission through the atmospheric limb (Tegtmeier et al., 2013; Sofieva et al., 2013).

16.3 GREENHOUSE GASES

All greenhouse gases are reported in dry mole fraction on the most recent scales (<https://www.esrl.noaa.gov/gmd/ccl/>). The scales are reviewed every two years at the WMO/IAEA Meetings on Carbon Dioxide, Other Greenhouse Gases and Related Measurement Techniques (GGMT) (WMO 2016c). The primary reference for greenhouse gases is a set of cylinders of natural air with known mole fractions of the studied gases. The primary scales are transferred to station working standards through secondary and tertiary gas standards in high-pressure cylinders.

16.3.1 Carbon dioxide (including $\Delta^{14}\text{C}$, $\delta^{13}\text{C}$ and $\delta^{18}\text{O}$ in CO_2 , and O_2/N_2 ratios)

Carbon dioxide is usually measured by active methods in the atmospheric boundary layer.

Most historical background atmospheric CO_2 measurements were made with non-dispersive infrared (NDIR) gas analysers, but a few programmes used a gas chromatographic method. Non-dispersive infrared instruments are based on the same principle that makes CO_2 a greenhouse gas: its ability to absorb infrared radiation. They measure the intensity of infrared radiation passing through a sample cell relative to radiation passing through a reference cell. It is not necessary to know the CO_2 mole fraction of the reference cell gas. Sample air is pumped from inlets mounted well away from the measurement building into the sample cell and for a fixed period of time measurements are taken. The sample is then flushed and replaced with reference or standard gas and further measurement gathered. The concentration of the CO_2 in the samples are determined by ratio comparison to the reference and/or standard gases. The GC method requires separation of CO_2 from other gases in the air sample, by reduction of CO_2 to CH_4 over a catalyst with H_2 . Detection of the CO_2 -derived CH_4 is achieved using a flame ionization detector. Chromatographic peak responses from samples are compared to those from standards with known CO_2 mole fractions to calculate the CO_2 mole fraction in the sample. Gas chromatography techniques are limited to a measurement frequency of one sample every few minutes.

Most of the newer methods for measurement of CO_2 use laser-based optical spectroscopic methods, like Fourier transform infrared (FTIR) absorption spectroscopy or high-finesse cavity absorp-

tion spectroscopy, which includes cavity ring-down spectroscopy (CRDS) and off-axis integrated cavity output spectroscopy (OA-ICOS). Among the advantages of these techniques are reduced calibration demands due to better linearity and stability.

Carbon dioxide abundances are reported in dry-air mole fraction, $\mu\text{mol mol}^{-1}$, often abbreviated as ppm, on the WMO CO₂ Mole Fraction Scale (WMO CO₂ X2007 scale, status as of 2018). Water vapour affects the measurement of CO₂ in two ways: (i) H₂O also absorbs infrared radiation and can interfere with the measurement of CO₂; (ii) H₂O occupies volume in the sample cell, while standards are dry. At warm, humid sites, 3 % of the total volume of air can be H₂O vapour. The impact of water vapour on the CO₂ measurement must therefore be considered. Drying to a dew-point of $-50\text{ }^{\circ}\text{C}$ is sufficient to eliminate interferences. The novel optical spectroscopic methods allow simultaneous determination of the H₂O vapour content, making it possible to correct for dilution due to H₂O and spectroscopic effects. Current best practice (see WMO, 2016c) recommends that water vapour must either be removed from the sample gas stream, or its influence on the mole fraction determination must be carefully quantified for each individual instrument.

An alternative method of CO₂ measurement that is generally applicable to many other trace gases is the collection of discrete air samples in vacuum-tight flasks. These flasks are returned to a central laboratory where the CO₂ mole fraction is determined by NDIR, GC or other techniques. This method is used where low-frequency sampling (for example, once a week) is adequate to define CO₂ spatial and temporal gradients, and for comparison with in-situ measurements as a quality-control step. This sampling strategy has the advantage that many species can be determined from the same sample.

Measurements of O₂/N₂ ratios and stable isotopes of CO₂ ($\delta^{13}\text{C}$ and $\delta^{18}\text{O}$) help to partition carbon sources and sinks between the ocean and biosphere. Isotopic measurements are often made from the same discrete samples used for CO₂ mole fraction measurements. No isotopic standards traceable to the International System of Units (SI) exist, but commonly agreed-upon reference material is maintained by the International Atomic Energy Agency (IAEA). The measurements are performed as part of the GAW CO₂ network.

A measurement method for stable isotope determination is isotope ratio mass spectrometry (IRMS), a specialization of mass spectrometry in which mass spectrometric methods are used to measure the relative abundance of isotopes in a given sample. The measurement set-up is described by the GAW Central Calibration Laboratory for stable isotopes at the Max Planck Institute for Biogeochemistry in Jena, Germany (http://www.bgc-jena.mpg.de/service/iso_gas_lab/pmwiki/pmwiki.php/IsoLab/Co2InAir). In recent years, optical analysers that report mole fractions of individual isotopologues have become increasingly available and are now in routine use. Many of these instruments can provide isotopic ratios with a repeatability of about 0.05 ‰ for $\delta^{13}\text{C}$ of atmospheric CO₂ and are valuable for continuous measurements. Unlike with mass spectrometric techniques, δ values from such instruments are often calculated from the ratio of individual measured mole fractions using tabulated absorption line strengths and are not from direct measurements of a standard material. Some corrections applicable to mass spectrometric methods, such as those for ¹⁷O and N₂O, are not required, but other corrections, such as for interference from other atmospheric components and instrument fluctuations, may be required depending on the method used to calculate the isotopic δ values from individual mole fractions. It is important to realize compatibility between the techniques before measurement results are made public.

Measurements of the changes in atmospheric O₂/N₂ ratio are useful for constraining sources and sinks of CO₂ and testing land and ocean biogeochemical models. The relative variations in O₂/N₂ ratio are very small but can now be observed by at least six analytical techniques. These techniques can be grouped into two categories: (i) those which measure O₂/N₂ ratios directly (mass spectrometry and gas chromatography), and (ii) those which effectively measure the O₂ mole fraction in dry air (interferometric, paramagnetic, fuel cell, vacuum-ultraviolet photometric). A convention has emerged to convert the raw measurement signals, regardless of technique, into equivalent changes in mole ratio of O₂ to N₂. For mole-fraction type measurements, this requires accounting for dilution due to variations in CO₂ and possibly other gases. If synthetic air is used as a reference material, corrections may also be needed for differences in Ar/N₂ ratio. There are currently about 10 laboratories measuring O₂/N₂ ratios worldwide. The O₂/N₂ reference is typically

tied to natural air delivered from high-pressure gas cylinders. As there is no common source of reference material, each laboratory employs its own reference. There is currently no CCL for O_2/N_2 . Hence it has not been straightforward to report measurements on a common scale, but several laboratories report results on a local implementation of the Scripps scale. There are no named versions yet.

The practice of basing O_2/N_2 measurements on natural air stored in high-pressure cylinders appears to be acceptable for measuring changes in background air, provided the cylinders are handled according to certain best practices, including orienting cylinders horizontally in order to minimize thermal and gravitational fractionation. Nevertheless, improved understanding of the source of variability of measured O_2/N_2 ratios delivered from high-pressure cylinders is an important need of the community. An independent need is the development of absolute standards for O_2/N_2 calibration scales to the level of 5 per meg or better.

Atmospheric $^{14}CO_2$ measurements are usually reported in $\Delta^{14}C$ notation, the per mil deviation from the absolute radiocarbon reference standard, corrected for isotopic fractionation and for radioactive decay since the time of collection. For atmospheric measurements of $\Delta^{14}C$ in CO_2 , two main sampling techniques are used: high-volume CO_2 absorption in basic solution or by molecular sieve, and whole-air flask sampling (typically 1.5–5 L flasks). Two methods of analysis are used: conventional radioactive counting and accelerator mass spectrometry. The current level of measurement uncertainty for $\Delta^{14}C$ in CO_2 is 2–5 ‰, with a few laboratories at slightly better than 2 ‰. Recommendations on calibration are provided in WMO (2016c).

Recommendations on quality assurance of CO_2 measurements (including $\Delta^{14}C$, $\delta^{13}C$ and $\delta^{18}O$ in CO_2 , and O_2/N_2 ratios) are reviewed every two years at the WMO/IAEA Meeting on Carbon Dioxide, Other Greenhouse Gases and Related Measurement Techniques. The report (WMO, 2016c) can be used as the most recent reference regarding calibration and measurement quality control.

16.3.2 Methane

Until recently, methane (CH_4) measurements were made almost exclusively using the GC-FID technique, and recommendations regarding the use of these systems can be found in the WMO/GAW measurement guideline (WMO, 2009). In the meantime, commercially available instruments based on spectroscopic techniques such as cavity ring-down spectroscopy (CRDS), off-axis integrated cavity output spectroscopy (OA-ICOS) and Fourier transform infrared (FTIR) spectroscopy have become more widely available and affordable. Spectroscopic techniques have been shown to have a number of advantages over the GC-FID method that will lead to improved accuracy of atmospheric CH_4 measurements (Zellweger et al., 2016). However, these techniques also require calibration (WMO, 2016c), and current best practice recommends that water vapour must either be removed from the sample gas stream, or its influence on the mole fraction determination must be carefully quantified for each individual instrument (Rella et al., 2013).

An alternative method of CH_4 measurement is flask sampling followed by off-line analysis, very similar to the approach described in 16.3.1 for CO_2 .

16.3.3 Nitrous oxide

Over the last few decades, measurements of atmospheric nitrous oxide (N_2O) have mainly been made using gas chromatographs with electron capture detectors (GC-ECD). However, this technique is very challenging for the detection of the small variations of N_2O in the troposphere. GC-ECDs are highly nonlinear and require a considerable amount of maintenance. Recommendations for N_2O measurements using GC-ECD are provided in the WMO/GAW measurement guideline WMO (2009), and recommendations on calibration and quality control are given in WMO (2016c).

Recently, spectroscopic analysers including high-finesse cavity absorption spectrometers with near-infrared laser sources, FTIR analysers and off-axis OA-ICOS analysers with mid-infrared laser sources became commercially available for N_2O measurements. These techniques normally exceed the performance of GC-ECD systems, but reaching the data quality objectives (WMO, 2016c) re-

mains challenging. A few field and laboratory studies (Lebeque et al., 2016; Vardag et al., 2014) show excellent performance and demonstrate the potential to replace the gas chromatographic techniques. However, further studies are needed to assess their long-term applicability and to identify optimum calibration strategies.

Collecting discrete samples of ambient air in flasks is an alternative method of monitoring N_2O . Flasks should be returned to a central laboratory for off-line analysis. Typical sampling frequencies are weekly or bi-weekly.

16.3.4 Halocarbons and SF_6

Halocarbons and SF_6 are usually measured quasi-continuously or from discrete air samples by active methods. Measurement guidelines for these species are not yet formalized in the GAW Programme though some guidance on calibration is provided by the World Calibration Centre for SF_6 (WMO, 2018).

SF_6 is typically measured using GC-ECD techniques on the same channel as N_2O . Analytical methods are described in WMO (2015e).

Global measurements of halocarbons are currently performed by the National Oceanic and Atmospheric Administration (NOAA) and the Advanced Global Atmospheric Gases Experiment (AGAGE). The measurement histories for both NOAA and AGAGE extend back to the late 1970s. Both groups measure halocarbons using GC-ECD and gas chromatography with mass spectrometry (GC-MS) techniques. Halocarbons measured include chlorofluorocarbons (CFCs), hydrochlorofluorocarbons (HCFCs), chlorinated solvents such as CCl_4 and CH_3CCl_3 , halons, hydrofluorocarbons (HFCs), methyl halides, NF_3 and SF_6 . For many halocarbons, measurement of mole fractions in the background troposphere requires sample pre-concentration. The AGAGE group operates a network of in-situ systems, while the NOAA group operates in-situ systems (for a limited number of gases) and a flask-based programme. For more information on instrumentation and sampling sites, see: <http://agage.eas.gatech.edu>, and <http://www.esrl.noaa.gov/gmd/hats/>. Reference gases are maintained by both AGAGE and NOAA and the individual scales are regularly compared.

16.3.5 Remote-sensing of greenhouse gases

There are several techniques used for remote-sensing of greenhouse gases. The Total Carbon Column Observing Network (TCCON, <https://tccon-wiki.caltech.edu/>) is a network of surface-based Fourier transform spectrometers recording direct solar spectra in the near-infrared spectral region. From these spectra, column-averaged abundances of CO_2 , CH_4 , N_2O , HF, CO, H_2O and HDO are retrieved. Observations in the mid-infrared (in the NDACC network, <http://www.acom.ucar.edu/irwg/>) allow for accurate measurements of column-averaged abundances of CH_4 , N_2O and CO.

16.4 REACTIVE GASES

The reactive gases considered in the GAW Programme include tropospheric ozone, carbon monoxide, volatile organic compounds, oxidized nitrogen compounds and sulphur dioxide. All of these compounds play a major role in the chemistry of the atmosphere and, as such, are heavily involved in interrelations between atmospheric chemistry and climate, either through control of ozone and the oxidizing capacity of the atmosphere, or through the formation of aerosols. The global coverage in terms of observations is entirely unsatisfactory for most of them, the only exceptions being surface ozone and carbon monoxide (Schultz et al., 2015).

Different reference standards and methods are used in the group of reactive gases. For more stable gases, the reference material can be prepared as a cylinder filled with air or another matrix with known gas mole fraction (for example, for CO, non-methane hydrocarbons and terpenes), while for others (such as ozone) only reference methods/instruments are possible.

16.4.1 Tropospheric (surface) ozone

Detailed measurement guidelines for measuring tropospheric ozone (surface ozone is a part of tropospheric ozone measured at the Earth's surface) are provided in WMO (2013).

The mole fraction most appropriate to the chemical and physical interpretation of ozone measurements is the mole fraction of ozone in dry air. However, ozone measurements are usually made without sample drying, because an efficient system for drying air and leaving the ozone content of the air unchanged has not been developed. It is recommended that ozone measurements be accompanied by measurements of water vapour mole fraction of sufficient precision that the ozone measurements could be converted to mole fractions with respect to dry air without loss of precision.

A number of techniques are used for measurements of ozone in the background atmosphere. These include:

- (a) Ultraviolet absorption techniques;
- (b) Chemiluminescence techniques;
- (c) Electrochemical techniques;
- (d) Cavity ring-down spectroscopy (CRDS) with NO titration;
- (e) Differential optical absorption spectroscopy (DOAS);
- (f) Multi-axis differential optical absorption spectroscopy (MAXDOAS);
- (g) Tropospheric ozone lidar.

Because of its high accuracy and precision, low detection limit, long-term stability, sufficient time resolution and ease of operation (almost no consumables), the UV absorption technique is recommended for use for routine surface ozone measurements at all GAW stations.

A review of measurement techniques, along with information on their applicability for use at GAW stations, is provided in WMO (2013). Note that only techniques a) to d) (those conducted in-situ) can be traceable via a chain of calibrations to the primary standard as recommended by GAW.

Differential optical absorption spectroscopy is a surface-based remote-sensing method suitable for observations of several trace substances. The instrument consists of a light source, a long ambient air open optical path generally between 100 m and several km, a retro-reflector and a spectrometer with a telescope, housed with the light source. The spectrometer observes the light source via the retro-reflector. The DOAS system uses Beer's law to determine the ozone concentration (averaged over the light path). In principle, DOAS should be a sensitive technique, but this is confounded by the inability of the system to regularly measure a definitive zero and determine the contribution of other UV absorbing gases and aerosols to the observed signal. DOAS may be used as an experimental technique (Platt and Stutz, 2008).

Multi-axis differential optical absorption spectroscopy is a surface-based remote-sensing method for observations of several trace substances. While this method is suitable for stratospheric monitoring, it is also possible to apply it for trace gas profile measurements in the upper and lower troposphere. However, since the retrieval procedures, as well as possible tropospheric interferences, are more complicated in the lower troposphere, it needs highly experienced personnel for extracting and calculating the mole fractions for the respective trace gases out of the various spectra. MAXDOAS measurements of ozone, nitrogen dioxide, formaldehyde, bromine monoxide (BrO) and other species are recommended especially for providing a link between surface-based and satellite measurements at selected GAW stations with extended research programmes (Hönninger et al., 2004).

Lidar (light detection and ranging) is a surface-based remote-sensing method for observations of several trace substances. For tropospheric ozone measurements, a lidar typically uses two or more wavelengths between 266 nm and 295 nm (Kuang et al., 2013). The chosen wavelengths are shorter than the ones used for stratospheric ozone detection (typically between 308 and 353 nm). Compared to in the stratosphere, higher ozone absorption efficiency is necessary in the troposphere in order to get enough sensitivity because of the lower ozone mixing ratios in the troposphere. Too much absorption means that most light is extinguished at lower elevations, making it difficult to collect measurement signals from higher elevations. The extreme dynamic range of the backscattering signal over the troposphere (some decades over a few kilometres of height) is a major technical problem. Lidar tropospheric ozone measurements are recommended especially for providing a link between surface-based and satellite measurements at selected GAW stations with extended research programmes.

16.4.2 Carbon monoxide

Detailed measurement guidelines for carbon monoxide measurements are provided in WMO (2010). The carbon monoxide calibration scale is evaluated every two years together with the scales of the major greenhouse gases WMO, 2016c). The most recent calibration scale can be found on the web page of the central calibration laboratory (<https://www.esrl.noaa.gov/gmd/ccl/>).

Measurements of CO are possible both in situ and by flask collection with subsequent analysis in the laboratory. In situ continuous observations provide information about CO variability on a time-scale ranging from seconds to one hour depending on the measurement technique. In contrast to flask sampling, continuous measurements allow for near-real-time data delivery.

In situ observations can be made using a broad variety of analytical techniques. Non-dispersive infrared radiometry is based on spectral absorption at 4.7 μm . It is frequently used for air pollution monitoring and occasionally also for continuous measurements at remote locations; however, instrument drift, limited precision and long averaging times are factors limiting the achievable data quality. Gas chromatography, when coupled with a number of different detectors (such as flame ionization (GC-FID) or hot mercuric oxide reduction/UV absorption (GC-HgO)) can provide high-precision and adequate detection limits. The HgO-reduction detector tends to have a non-linear response over the range of atmospheric CO, and requires careful, repeated multipoint characterization of the detector response. The GC-FID technique requires catalytic conversion of CO to CH₄. For confidence in the results, the catalytic conversion efficiency must be determined on a regular basis making the proper maintenance, instrument calibration and provision of accurate measurements challenging. Gas chromatography measurements are quasi-continuous in nature and therefore may not detect fast changes of mole fractions that can be captured by high-frequency measurements.

In recent years, many alternative techniques have become available. These include resonance fluorescence of CO (induced by a high-frequency discharge) in the vacuum ultraviolet (VURF), spectroscopic techniques based on CRDS, quantum cascade laser (QCL) spectroscopy, and Fourier transform infrared (FTIR) absorption spectroscopy. The CRDS technique operates using lasers in the near-infrared and was previously mainly used for measurements of carbon dioxide, methane and ammonia. The QCL technique measures in the mid-infrared, and commercial instruments are available that can determine both CO and N₂O with a single laser. An overview of the performance of most analytical techniques can be found in Zellweger et al. (2009, 2012).

Remote-sensing of the CO column from the ground is undertaken by the TCCON network (<http://www.tcon.caltech.edu/>) using surface-based Fourier transform spectrometers in the near-infrared spectral region (Wunch et al., 2010).

16.4.3 Volatile organic compounds

The measurement of VOCs is complex due to the many different molecules present in the atmosphere. While systematic surveying of many of these species is important for air quality purposes, the low concentrations of VOCs away from their sources imply that only a few molecules can be measured routinely in the background atmosphere. A core set of molecules recommended for

measurement in the GAW Programme with suggested measurement methods is provided in Table 16.1.

The measurement guidelines for VOCs are currently under development in collaboration with the ACTRIS Research Infrastructure. A standard operating procedure for taking air samples with stainless steel canisters is available in WMO (2012). General recommendations on VOC measurements can be found in WMO (2007). Regular GAW VOC workshops review the status of VOC measurements in the GAW Programme (http://instaar.colorado.edu/arl/GAW_VOC_meeting.html) and provide further guidance on the development of measurement techniques, quality assurance and gas standards.

Table 16.1. List of the priority VOCs in the GAW Programme

TABLE: Table horizontal lines						
Molecule	Lifetime (assuming OH concentration is 10^6 cm^{-3})	Importance to GAW	Steel flask ^a	Glass flask	Analysis method ^b	
1. Ethane	1.5 months	<ul style="list-style-type: none"> - Tracer for fossil fuel emissions - Natural sources - Biomass burning - Fossil fuel - Trend in size of seasonal cycle - Indicator of halogen chemistry 	✓	✓	GC-FID	
2. Propane	11 days	<ul style="list-style-type: none"> - Tracer for fossil fuel emissions - Biomass burning 	✓	✓	GC-FID	
3. Acetylene	15 days	<ul style="list-style-type: none"> - Motor vehicle tracer - Biomass burning tracer - Ratios to the other hydrocarbons 	✓	✓	GC-FID	
4. Isoprene	3 hours	<ul style="list-style-type: none"> - Biosphere product - Sensitive to temperature/land - Used for climate change - O₃ precursor - Oxidizing capacity - Precursor to formaldehyde 	?	?	GC-FID PTR-MS	
5. Formaldehyde	1 day	<ul style="list-style-type: none"> - Indicator of isoprene oxidation - Biomass burning - Comparison with satellites- 	-	-	DOAS	
6. Terpenes	1-5 hours	<ul style="list-style-type: none"> - Precursors to organic aerosols 	-	-	GC-MS PTR-MS	
7. Acetonitrile	0.5-1 year	<ul style="list-style-type: none"> - Biomass burning indicator - Biofuel burning indicator 	-	?	GC-MS PTR-MS	

8. Methanol	12 days	- Sources in the biosphere (methane oxidation) - Abundant oxidation product	-	?	GC-FID PTR-MS
9. Ethanol	4 days	- Tracer of alternative fuel usage	-	?	GC-FID PTR-MS
10. Acetone	1.7 months	- Abundant oxidation product - Free radical source in the upper troposphere	?	?	GC-FID PTR-MS
11. Dimethyl sulphide	2 days	- Major natural sulphur source - Sulphate aerosol precursor - Tracer of marine bioproductivity	?	?	GC-FID PTR-MS
12. Benzene	10 days	- Tracer of combustion - Biomass burning indicator	✓	?	GC-FID GC-MS
13. Toluene	2 days	- Ratio to benzene used for air mass age - Precursor to particulates	-	?	GC-FID GC-MS
14. Iso/normal butane	5 days	- Chemical processing indicator - Lifetime/ozone production	✓	✓	GC-FID GC-MS
15. Iso/normal pentane	3 days	- Ratio provides impact of NO ₃ chemistry and oil and natural gas sources	✓	✓	GC-FID GC-MS

Notes:

a "✓" indicates state of current practice, "-" stands for not measured in flasks (only on-line measurements), while "?" means that stability in flasks is not well known

b GC-FID = gas chromatography flame ionization detection; GC-MS = gas chromatography mass spectrometry; DOAS = differential optical absorption spectroscopy; PTR-MS = proton transfer reaction mass spectrometry

Measurements of low molecular weight aliphatic and aromatic hydrocarbons (C₂-C₉) have been made successfully for many years, predominantly in short-term regional experiments. The preferred analytical method for these compounds, which include the molecules 1-4 and 12-15 of Table 16.1, is GC-FID. Air samples, from flasks or in situ, are normally pre-concentrated using cryogenic methods or solid adsorbents. An alternative technique is GC-MS. Although GC-MS is potentially the more sensitive method, it is typically subject to greater analytical uncertainties (changes in instrument response over time, detection of common, low-mass fragments). However, GC-MS may be valuable for the detection of certain hydrocarbons in very remote locations where ambient levels may be below the detection limit of a typical GC-FID.

The recommended analytical technique for monoterpenes is GC-MS. Although it is possible to measure some terpenes using a FID, the complexity of the chromatographic analysis (co-eluting peaks, particularly with aromatics) makes peak identification and quantification difficult. The GC-MS method gives better sensitivity.

Oxygenated hydrocarbons, including the target compounds 8-10 (Table 16.1), can also be measured using GC-FID or GC-MS. Particular care should be taken with sample preparation (including water removal), and inlet systems must be designed to minimize artefacts and component losses commonly encountered with oxygenate analysis. Acetone and methanol can also be measured using proton transfer reaction mass spectrometry (PTR-MS). An advantage of PTR-MS is that it is an online method that does not require the pre-concentration of samples. However, it is less sensitive than GC methods, and there are potential interferences from isobaric compounds, such as

O_2H^+ and methanol. As the stability of oxygenated VOCs in grab samples (stainless steel or glass flasks) remains highly uncertain, it is suggested that these species be measured primarily by online methods at a selection of surface-based measurement stations. The successful storage of acetone in certain flasks has been reported, so the possibility of analysing this compound in the glass or stainless steel flask network should be investigated.

Formaldehyde (HCHO) is not stable in flasks and has to be measured in situ. Methods of analysis include the Hantzsch fluorometric (wet chemical) method (Nash, 1953) or DOAS. Both are relatively complex and would require specialist training for potential operators. It is unlikely, therefore, to be able to make measurements at more than a few ground stations. Formaldehyde is routinely detected by satellites. Satellite retrievals yield total vertical column abundances, and an important objective of the GAW Programme would be to provide periodic surface-based measurements at selected sites for comparison/calibration purposes (ground truthing).

The feasibility of HCHO measurements with PTR-MS (Wisthaler et al., 2008; Warneke et al., 2011) and QCL (Herndon et al., 2007) was shown during limited measurement campaigns. Their applicability for long-term routine HCHO measurements has not yet been tested.

Acetonitrile is preferably measured with GC-MS, because this compound is relatively insensitive to FID detection. Measurements of acetonitrile have also been reported using various reduced gas and nitrogen-specific detectors. Many recently reported atmospheric measurements of acetonitrile have been made using PTR-MS or atmospheric pressure chemical ionization mass spectrometry (AP-CIMS). The stability of acetonitrile in grab samples is highly uncertain, so grab sampling is not acceptable in the framework of GAW and measurements may be limited to a few selected comprehensive measurement sites.

Dimethyl sulphide (DMS) can be measured by GC-FID, gas chromatography using a flame photometric detector (GC-FPD), GC-MS and PTR-MS. However, as DMS concentrations can be measured routinely as part of a standard non-methane hydrocarbon (NMHC) analysis, GC-FID analysis of the whole air samples would be the simplest choice of measurement strategy. There is evidence in the literature that DMS is stable in some flasks, so its measurement as a component of a flask network is quite feasible. It is also desirable to make in situ measurements of DMS at least in the early stage of operation of the flask network to ensure method compatibility.

16.4.4 Nitrogen oxide

The sum of nitric oxide (NO) and nitrogen dioxide (NO_2) has traditionally been called NO_x . The sum of all nitrogen oxides with an oxidation number greater than 1 is called NO_y . Their measurement in the global atmosphere is very important since NO has a large influence on both ozone and the hydroxyl radical (OH). NO_2 is now being measured globally from satellites, and these measurements suggest that substantial concentrations of this gas are present over most of the continents. A large reservoir of fixed nitrogen is present in the atmosphere as NO_y . The influence of the deposition of this reservoir on the biosphere is not well known at present but could be substantial.

Nitrogen oxide (NO and NO_2) measurements can be done by passive, active and remote-sensing techniques. The active techniques can be divided into time-integrating and in situ techniques: time-integrating techniques consist of a sampling step usually involving liquid-phase sample collection and offline analysis, whereas in situ (continuous) measurements directly analyse the sample air. Passive methods are always time-integrating. Active time-integrating methods comprise the Saltzman method and related methods like the Griess or sodium iodide method. The latter is being used, for example, in the Cooperative Programme for Monitoring and Evaluation of the Long-range Transmission of Air Pollutants in Europe (EMEP, <http://www.emep.int>) network. Due to the high reactivity of NO_x , flask sampling is impossible.

Ozone-induced chemiluminescence detection is the most widely used among the in situ techniques. These instruments are typically very sensitive to NO; however, they cannot measure NO_2 . Thus, NO_2 must be converted to NO before detection. The instrument makes measurements in a NO mode and then a NO + NO_2 mode. The difference, when conversion efficiency is determined carefully, gives the NO_2 mixing ratio. Thus, a high time resolution (< 10 min) is recommended to ensure sampling of the same air mass during subsequent NO and NO_x measurements. The conver-

sion of NO_2 to NO is achieved by photolysis of NO_2 at wavelengths $320 < \lambda < 420$ nm using a photolytic converter (PLC) with an arc lamp or a blue-light converter (BLC) with light-emitting diodes (LEDs). Advantages of LEDs are the substantially longer lifetime and nearly constant conversion efficiencies, the mechanical simplicity and the simple on/off characteristic of the LED (no additional valves/dead volumes). Depending on the chosen LED, the conversion efficiency can be smaller or even larger than with PLC. However, the spectral range of the new LED-based converters and temperature effects need to be considered and characterized in order to prevent artifacts due to nitrous acid (HONO) and peroxyacetyl nitrate (PAN). With careful selection, the use of UV-LED converters is recommended for GAW NO_2 measurements.

The use of molybdenum converters for NO_2 to NO conversion is strictly discouraged, as this conversion technique is not selective of NO_2 but also converts other oxidized nitrogen species in different quantities. Already-existing measurements with Mo converters should be marked as $\text{NO}_{2(\text{Mo})}$ or NO_{2+} .

Detailed measurement guidelines for reactive nitrogen measurements are currently being developed in collaboration with the ACTRIS Research Infrastructure. The focus there is mostly on NO and NO_2 point measurements because their measurements are presently more extensive and robust and allow for implementation of a complete quality assurance system. Recommendations on NO and NO_2 measurements can be found in WMO (2017b).

The luminol-CLD method (Kelly et al., 1990) measures NO_2 directly and NO indirectly after oxidation. Since the sensitivity depends strongly on the quality of the luminol solution, which decreases during use due to ageing, frequent re-calibration is needed.

In addition to these methods, optical absorption techniques for NO_2 detection have been developed, including tunable diode laser absorption spectroscopy (TDLAS), differential optical absorption spectroscopy, laser-induced fluorescence (LIF), Fourier transform infrared absorption spectroscopy and cavity ring-down spectroscopy. They all measure NO_2 directly. Recent developments in CRDS for the measurement of NO_2 and of NO as NO_2 after oxidation by ozone show some promise, but the measurements still suffer from uncertainties in the zero level.

Recently, the suitability of research-type quantum cascade laser instrumentation for continuous and direct measurements of NO and NO_2 was shown (Tuzson et al., 2013). This technique may become an alternative standard method in the future.

Also recently, cavity attenuated phase shift (CAPS) monitors have become commercially available. A side-by-side intercomparison experiment at ACTRIS showed excellent results. However, a current lower detection limit (LDL) of few tens of ppt make this technique appropriate for either rural or anthropogenic-influenced sites (Ge et al., 2013) but not fully suitable for remote ones with typical NO_2 mole fractions below 50 ppt. Moreover, calibration of NO_2 specific detectors remains challenging as reference gases with ambient levels of NO_2 in high pressure cylinders are known to be susceptible to instabilities.

At present, there is no mature point measurement technique that can compete with the ozone-induced chemiluminescence detection of NO at remote locations. Passive and active time-integrating methods are not accepted in the GAW Programme due to their poor selectivity and time resolution.

More details on procedures for standard operations can be found at :
<http://actris.nilu.no/Content/?pageid=68159644c2c04d648ce41536297f5b93>.

Ground-based remote sensing techniques for NO_2 are currently being developed towards more operational use and better defined uncertainties in the framework of ACTRIS. MAXDOAS (Multi-Axis DOAS) instruments are increasingly used for vertical profile measurements, FTIR and zenith-sky UVVIS measurements are also used for NO_2 vertical column information.

16.4.5 Sulphur dioxide

Measurement guidelines for SO₂ observations are so far not available in the GAW Programme. General recommendations are given in WMO (2001). However, the GAW Scientific Advisory Group for Reactive Gases is planning to establish the same quality assurance system for SO₂ (including measurement guidelines and central facilities) as it is doing at present for nitrogen oxides after the latter is completed.

There are various measurement techniques for determining atmospheric SO₂. EMEP is using integrating techniques, such as an alkaline-impregnated filter (pack) or coated annular denuder, both followed by ion chromatography in a central laboratory. These methods yield potentially more accurate results, but with a time resolution of usually one sample per day, as is typical for integrative techniques. Additionally, they require frequent attention, and personnel costs for filter analysis are high.

In the group of in-situ measurements, the TCM (photometry after reaction of SO₂ with tetrachloromercurate) and the pulsed fluorescence methods are widely used. TCM has a high accuracy but high low detection limit (LDL), and the handling of mercury in the laboratory could be harmful. Even though the response of the pulsed-fluorescence sensor is slower, its ease of calibration, dependability, accuracy and SO₂ specificity make it preferable. More sensitive gas chromatographic techniques are also available. However, they require significant technical expertise and regular attention. In order to enhance the sensitivity, some fluorescence analysers are equipped with more selective excitation filters. For example, two mirror assemblies are connected in series and specially selected PMTs are employed. Numerical corrections for interfering substances are possible; however, this is not necessary in rural or remote areas. The typical LDL which can be reached with these provisions is some 50 ppt. A further enhancement of sensitivity may be possible by means of a second channel in which only SO₂ is removed, leading to a highly specific read-out after subtraction of both channels.

Since SO₂ has a short atmospheric lifetime, understanding the sulphur cycle requires knowledge of the source and sink terms. This is best accomplished with sampling frequencies of less than 1 h. Therefore, the best technique for long-term monitoring of SO₂ today is a combination of the pulsed-fluorescence analyser and filter sampling. Filter samples should be exposed at intervals, but often enough to act as a quality control for the continuous analyser.

More detailed recommendations can be found in EMEP (2014), available at <http://www.nilu.no/projects/ccc/manual/index.html>.

16.4.6 Molecular hydrogen

Detailed measurement guidelines for molecular hydrogen are currently not available in the GAW Programme.

Molecular hydrogen is reported as dry mole fraction on the most recent scale (WMO, 2016c). Measurements of H₂ are possible both in situ and by flask collection with subsequent analysis in the laboratory. An example of the measurement system set up at the GAW global stations is given in (Grant et al., 2010).

Molecular hydrogen measurements are performed with gas chromatography followed by hot mercuric oxide reduction/UV absorption detection. An alternative GC set-up with pulsed-discharge detectors (PDD) has a more linear detector response and provides better repeatability for molecular hydrogen measurements (Novelli et al., 2009).

Problems with instability of H₂ in reference gases have frequently been experienced (Jordan and Steinberg, 2011). Therefore, recommendations on calibration and quality assurance of H₂ measurements available in WMO (2016c) should be consulted.

16.5 ATMOSPHERIC WET DEPOSITION

Atmospheric wet deposition refers to the gases and particles deposited by precipitation on the Earth's surface. These gases and particles have a wide variety of sources and compositions and generally are present in trace amounts in the atmosphere and in precipitation. These trace materials are captured by precipitation as it forms in the atmosphere and falls to the Earth. The deposited materials constitute an important contribution to the mass balance of pollutants associated with long-range transport. These materials not only affect the chemistry of precipitation but also can affect the chemistry of the terrestrial and aquatic surfaces on which they are deposited. The effects can be either harmful or beneficial, and they can be either direct or indirect. For example, acidic wet deposition is an environmental problem that results from combustion of fossil fuels. It occurs when oxides of sulphur and nitrogen, emitted during combustion, are transformed in the atmosphere and become acidic sulphate and nitrate in precipitation. Other trace materials that occur in wet deposition include sea salt, nutrients, chemicals found in soil particles, toxic organic and inorganic chemicals, organic acids, etc. Research has shown that some wet-deposited chemicals can stimulate marine biotic production, potentially linking atmospheric wet deposition to the carbon cycle and climate change.

Measuring the chemistry of precipitation tells us what trace materials are present in wet deposition and in what amounts. This information can be used to evaluate air quality and to identify and track changes in gaseous and particulate emissions to the atmosphere. In short, precipitation chemistry measurements provide information on the exchange of trace materials between the atmosphere and the land/oceans, and hence are important in furthering our understanding of the chemical cycles of these materials, especially those that can result in damage to terrestrial and aquatic systems or affect our climate.

Special care is required when planning precipitation chemistry measurements to ensure that they are representative. Though the measurements are made at a particular location, on average they should represent measurements in the surrounding region. In general, the sample collection site should be characteristic of the land use in the region. For example, the site in an area dominated by agricultural activities should have an agricultural setting. This quality of spatial representativeness should extend across seasons and even over years. Ideally, a site would be both spatially and temporally representative. Contamination of a localized nature from agricultural, industrial or other human activities must be avoided, as must the local impact of natural sources, such as oceanic shores, volcanoes or fumaroles. Sample collection should not be impacted by trees or other vegetation, and the on-site topography should be level and the exposure relatively unaffected by wind patterns that may result in an unrepresentative catch of rain and snow. Human contact with the sample or contact with anything that might change the sample chemistry must be avoided as well. Ensuring representative precipitation chemistry measurements entails strict adherence to requirements for site location, site conditions, equipment installation and site operational protocols and maintenance. These requirements are documented in WMO (2004a).

Precipitation chemistry monitoring can be divided into sample collection activities and chemical analysis activities.

16.5.1 Sample collection

The primary goal of the GAW Precipitation Chemistry Programme is to collect wet-only deposition samples. This means that the samplers are exposed only during precipitation and trace materials in the samples are deposited only by precipitation. The trace materials from dust or fine particles or gases deposited during dry weather are excluded. This makes it possible to study precipitation chemistry without contamination from dry deposition. More importantly, the equipment and methods for collecting a representative wet deposition sample are inappropriate for collecting a representative dry deposition sample. The physical and chemical processes affecting wet and dry deposition are distinctly different.

The best way to ensure collection of a wet-only sample is to employ an automated sampler that is open only during precipitation. A typical automated, wet-only deposition sampler has the following components: a precipitation sample container (funnel-and-bottle, bucket, etc.), a lid that opens and closes over the sample container orifice, a precipitation sensor, a motorized drive mechanism

with associated electronic controls and a support structure to house the components. The containers should have sufficient volume to hold all precipitation collected during the sampling period. A system that can be activated manually for testing, cleaning and routine maintenance is recommended. A modular design that allows removal of individual components, such as the sensor, facilitates rapid repair with a minimum of tools and expertise. An alternative to using an automated sampler is to collect samples by manually exposing a sample container at the very onset of precipitation and closing it as soon as precipitation ceases. This requires diligent round-the-clock observers alert to weather conditions; as a consequence, manual sample collection is very labour-intensive.

To complement the collection of wet-only deposition samples, the GAW Precipitation Chemistry Programme requires every site to measure precipitation depths using the standard precipitation gauge designated by National Meteorological and Hydrological Services or its equivalent (see Volume I, Chapter 6 of this Guide). Manual gauges are preferred. Precipitation depths are used to calculate the mass of a chemical deposited by precipitation on an area of the Earth's surface (called the wet deposition flux or loading). Standard precipitation gauges are designed to be the most accurate and representative means of measuring precipitation depths. Thus, each site must operate a precipitation gauge in parallel with its precipitation chemistry sampler. Precipitation chemistry sampler volumes are used to calculate wet deposition fluxes only when the standard gauge fails or is temporarily out of service. The data record should document such cases.

The highest priority of the GAW Precipitation Chemistry Programme is to collect a wet-only sample on a daily (24 h) basis with sample removal set at a fixed time each day, preferably 0900 local time. Should the resources be inadequate to collect and analyse daily samples, multi-day sampling periods up to one week is the next highest priority. Alternative sampling protocols are described in WMO (2004a). Collecting samples daily reduces the potential for the degradation of labile chemical species and for other physical and chemical changes in the sample while it is held in the field sampler. Not only is the sample integrity less likely to be compromised by a daily sampling protocol but the data have greater value as well. Storm trajectory analyses and source-receptor models are much less complicated when precipitation is more likely to have come from a single event or storm. Multi-day and one-week samples are much more likely to contain precipitation from several storms, each occurring under different meteorological settings. Further, daily data can be integrated mathematically to determine weekly or longer-term averages, but weekly data cannot be differentiated into daily components without making substantial assumptions.

Containers used to collect, store and ship samples should be unbreakable and sealable against leakage of liquids or gases. High-density polyethylene containers are recommended. All sample containers must be cleaned with deionized water of known and assured quality. The report (WMO, 2004a) contains detailed descriptions of the procedures for cleaning containers and ensuring that cleanliness standards are maintained throughout the collection, storage and shipment of samples.

16.5.2 Chemical analysis

The following chemical parameters are recommended for analysis in GAW precipitation samples: pH, conductivity, sulphate, nitrate, chloride, ammonium, sodium, potassium, magnesium and calcium. Analyses for formate and acetate are recommended for areas suspected of having high organic acid concentrations. Nitrite, phosphate and fluoride concentrations also may be important in certain areas, although their analyses are not required by GAW at this time. Preferred analytical methods are given in Table 16.2.

ELEMENT 1: Floating object (Bottom)

Table 16.2. Chemical parameters required for analysis in the GAW Precipitation Chemistry Programme with recommended analytical methods

TABLE: Table horizontal lines

<i>Analyte</i>	<i>Status</i>	<i>Preferred methods^a</i>
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pH	Required	Glass electrode
Conductivity	Required	Conductivity cell
Alkalinity	Optional	Titration
Cl ⁻ , NO ₃ ⁻ , SO ₄ ²⁻	Required	IC
NH ₄ ⁺	Required	IC, AC
Na ⁺ , K ⁺ , Ca ²⁺ , Mg ²⁺	Required	IC, ICP, AAS/AES
Organic acids ^b	Optional	IEC, IC
NO ₂ ⁻ , F ⁻	Optional	IC
PO ₄ ³⁻	Optional	IC, AC

Notes:

- a IC = ion chromatography; AC = automated colorimetry; ICP = inductively coupled plasma spectrometry; IEC = ion exclusion chromatography; AAS = atomic absorption spectrometry; AES = atomic emission spectrometry
- b For areas with high organic acid concentrations, formate and acetate analyses are recommended.

END ELEMENT

Past experience from regional networks and laboratory intercomparisons has shown that measuring pH in precipitation is difficult due mainly to the low ionic strength of the samples. Samples may also degrade due to biological activities and should therefore be kept refrigerated until the time of analysis, when they are brought to room temperature. The pH measurements should be carried out within two days of sample arrival in the laboratory.

Commercial pH meters are available with different specifications and options. A pH meter should have both an intercept and slope adjustment and should be capable of measuring to within ± 0.01 pH unit. Combination electrodes containing both measuring and reference functions are often preferred since they require smaller amounts of a sample, but a set of two electrodes may also be used with the pH meter. The measuring glass electrode is sensitive to hydrogen ions and the reference electrode can be calomel, or silver/silver chloride. Low ionic strength electrodes are now available commercially. Other reference electrodes can also be used as long as they have a constant potential. When selecting any electrode, confirm its ability to measure low ionic strength solutions by measuring a certified reference material. Response time should be less than 1 min and the addition of potassium chloride (KCl) should not be needed.

The conductivity of a solution is the reciprocal value of its specific resistance and can be directly measured using a conductivity bridge with a measuring cell. Conductivity varies with the temperature of the solution and is proportional to the concentration and the species of free ions present in the solution. Since the conductivity also depends on the electrode area and its spacing, the measuring apparatus has to be calibrated to obtain the cell constant or to adjust the meter. A KCl solution of known concentration and conductivity is used for calibration. Conductivity is measured and expressed in units of microsiemens per centimetre ($\mu\text{S cm}^{-1}$), corrected to 25 °C. The conductivity range of precipitation samples is 5 to 1 000 $\mu\text{S cm}^{-1}$. In case of small sample volumes, the aliquot that is used for conductivity measurement can be used for pH determination. If this is done, the conductivity should be measured first to avoid any possible error due to salt contamination from the pH electrode.

The apparatus for conductivity measurements consists of:

- (a) A conductivity meter (with operating range of 0.1 to 1 000 $\mu\text{S cm}^{-1}$; or, better, 0.01 to 1 000 $\mu\text{S cm}^{-1}$). Precision has to be within 0.5 % of the range and accuracy at 1 % of the range;

- (b) A conductivity cell (if the values in precipitation samples are expected to be mainly very low ($<20 \mu\text{S cm}^{-1}$), use special conductivity cells, with a low cell constant);
- (c) A thermometer (0 °C to 40 °C / 0.1 °C);
- (d) A water bath at 25 °C;
- (e) A polyethylene or glass vessel corresponding to the diameter of the cell used.

Ion chromatography (IC) has been widely used in recent years to analyse major anions and cations in precipitation, mainly in combination with electrochemical detection.

Sulphate, nitrate, chloride as well as other anions in precipitation are separated on an ion exchange column because of their different affinities for the exchange material. The material commonly used for anion separation is a polymer coated with quaternary ammonium active sites. After separation, the anions pass through a suppressor that exchanges all cations for H^+ ions. Instead of strong acid cation exchange columns, today micro membrane and self-regenerating suppressors with chemical or electrochemical regeneration are used. As a result of the suppression reaction, corresponding acids of the eluent ions and of chloride, nitrate and sulphate will reach the conductivity detector. A decreased basic conductivity and higher analytical signals now allow the detection of anions in the lower $\mu\text{g L}^{-1}$ range.

There are several anion exchangers with different properties available on the market. The time for one analysis and the quality of separation of single signals are dependent on the type of column and eluent, and on the concentration and flow rate of the eluent.

Any anions with a retention time similar to that of the main anions in the solution can cause interference. For example, when NO_2^- is present, it elutes just after Cl^- , which can cause the peak to be asymmetric. In rare cases, when the concentration of Cl^- is very high compared with that of NO_3^- , it can also influence the determination of NO_3^- . The manual should be consulted to see how different integration programmes handle this problem.

With care, up to several thousand analyses can be performed with the same anion separator column. The most effective method of protecting the separator column is to use a pre-column in front of it. Details are provided by the manufacturers in the manuals for the columns.

The principle of cation measurements is the same as that of anion determination except that different column materials are used and the suppressor column is often omitted. The material commonly used for cation separation is a cation exchange resin with active surface groups. Sodium, ammonium, potassium, calcium and magnesium ions are detected by a conductivity detector, without changing the eluent when certain columns are used. In other columns, monovalent cations (Na^+ , NH_4^+ , K^+) are determined using one eluent and divalent cations (Mg^{2+} and Ca^{2+}) with another eluent (because of their higher affinity to the resin).

Any cation with a retention time similar to that of the main cations may cause interference. For example, in samples with high concentrations of Na^+ , the peak of NH_4^+ becomes asymmetrical and often causes significant error. In this case, measurement using more dilute eluent could improve the separation of peaks.

Sodium, potassium, magnesium and calcium in precipitation are often analysed by atomic spectroscopic methods. Both flame (atomic absorption spectrometry (AAS) and atomic emission spectrometry (AES)) and plasma (inductively coupled plasma atomic emission spectrometry (ICP-AES) and inductively coupled plasma mass spectrometry (ICP-MS)) based methods can be used. For these ions, ion chromatography has no special advantage in terms of sensitivity, precision and accuracy over the spectroscopic methods, although analysis of all ions in one sample run is not possible with flame AAS or AES (single element methods).

The ions in the sample solution are transformed to neutral atoms in an air/acetylene flame. Light from a hollow cathode or an electrodeless discharge lamp (EDL) is passed through the flame. In the AAS mode, light absorption of the atoms in the flame is measured by a detector following a

monochromator set at the appropriate wavelength. Light absorption is proportional to the ion concentration in the sample. In the AES mode, the light emitted from the atoms excited in the flame is measured. Most commercial instruments can be run in both modes. Atomic emission spectrometry is the preferred mode for sodium measurements.

In atomic absorption spectrometry, both ionization and chemical interferences may occur. These interferences are caused by other ions in the sample, which reduce the number of neutral atoms in the flame. Ionization interference is avoided by adding a relatively high amount of an easily ionized element to the samples and calibration solutions. For the determination of sodium and potassium, caesium is added. For the elimination of chemical interferences from aluminium and phosphate, lanthanum can be added to the samples and calibration solutions for calcium and magnesium.

Formic and acetic acids (HCOOH and CH_3COOH , respectively) are major chemical constituents of precipitation in both continental and marine regions. Available evidence suggests that these compounds originate primarily from natural biogenic sources; both direct emissions (over continents) and emissions of precursor compounds appear to be important. Biomass and fossil fuel combustion also result in the emission of carboxylic acids and/or their precursors to the atmosphere.

Carboxylic acids in precipitation are very unstable and rapidly disappear from unpreserved samples. To generate reliable data, precipitation must be sampled on a daily or event basis and immediately preserved with the addition of a biocide such as chloroform (CHCl_3). Typically, 250 ml aliquots of sample (or less for low volume events) are treated with 0.5 ml of CHCl_3 . Samples are then tightly sealed and refrigerated until analysis.

Carboxylic species can be analysed by both IC (using a dilute eluent) and ion exclusion chromatography (IEC). However, acetate and propionate typically co-elute when analysed by IC and are thus impossible to resolve quantitatively. The IEC method exhibits fewer interferences associated with co-eluting species and is thus preferred for analysis of precipitation samples.

For analysis by IEC, samples are added to a hydrochloric acid (HCl) eluent which then flows through a separator column, a suppressor column and a detector. Resin in the separator column partitions anions using the principle of Donnan exclusion; anions are retained and sequentially separated based on their respective $\text{p}K_{\text{a}}$ s and van der Waals interactions with the resin. Anions of stronger acids with lower $\text{p}K_{\text{a}}$ s, such as H_2SO_4 , HNO_3 and HCl , are effectively excluded and co-elute early in the chromatogram; those of weaker acids with higher $\text{p}K_{\text{a}}$ s, such as HCOOH and CH_3COOH , elute later in the chromatogram. The suppressor column incorporates a cation exchange resin with silver added to the exchange sites; H^+ exchanges with the silver; the released silver subsequently reacts with Cl^- in the eluent to form silver chloride (AgCl), which precipitates within the column. Acid analytes exit the suppressor in a stream of deionized water. Detection is by conductivity.

16.6 AEROSOLS

Atmospheric aerosols are important for a diverse range of issues including global climate change, acidification, regional and local-scale air quality, and human health. The climate impact of aerosols is a result of direct radiative effects and indirect effects on cloud properties. Regional problems include potential impacts on human health and mortality and environmental impacts such as visibility impairment. Major sources of aerosols include urban/industrial emissions, smoke from biomass burning, secondary formation from gaseous aerosol precursors, sea salt and dust. Outstanding problems include determining the natural sources of aerosols and the organic fraction.

Table 16.3 provides a list of aerosol parameters recommended for measurement in the GAW Programme. Comprehensive measurement guidelines for aerosol measurements are provided in WMO (2016d) and WMO (2011).

Table 16.3. List of comprehensive aerosol variables recommended for long-term measurements in the global network**TABLE: Table horizontal lines**

<i>Variable</i>	<i>Frequency of observation</i>
Multiwavelength aerosol optical depth	Continuous
Mass concentration in two size fractions (fine, coarse)	Continuous
Mass concentration of major chemical components in two size fractions	Continuous
Light absorption coefficient at various wavelengths	Continuous
Light scattering and hemispheric backscattering coefficient at various wavelengths	Continuous
Aerosol number concentration	Continuous
Aerosol number size distribution	Continuous
Cloud condensation nuclei number concentration at various supersaturations	Continuous
Vertical distribution of aerosol backscattering and extinction	Continuous
Detailed size fractionated chemical composition	Intermittent
Dependence of aerosol variables on relative humidity, especially aerosol number size distribution and light scattering coefficient	Intermittent

Source: WMO (2016d)

16.6.1 Aerosol chemical measurements

At present, filter collection of ambient aerosols, followed by laboratory analyses, still remains the most commonly used and cost-efficient method available for the determination of aerosol chemical composition despite well-documented artefacts. Artefacts are very often linked to the presence of semi-volatile species that can either condense upon sampling (positive artefact) or evaporate from the filter media after sampling (negative artefact). A number of methods have been proposed to account for these artefacts (Cavalli et al., 2010).

The optimal set-up for the characterization of chemical properties of aerosol would be composed of a series of denuders to remove condensable species present in the gas phase (limiting positive artefacts) and a filter pack collecting both particles and condensable species re-emitted from the first filter (accounting for negative artefacts). Filter packs have been developed, consisting of a sandwich of filters and collection media of various types in series, to collect aerosols and selectively trap gases and aerosol volatilization products. Ideally, sampling for inorganic and carbonaceous species is performed with two different sampling lines since different kinds of denuders and filter media are required for analyses of elemental and organic carbon (EC/OC), and inorganic species. A third line could be implemented for sampling and analysis of elemental aerosol composition. The optimal set-up for OC/EC sampling is described in EN-16909 (CEN, 2017) and in (Cavalli et al., 2010).

Clearly, methods for artefact limitation add complexities with respect to plain filter measurements. For GAW purposes, considering remoteness and the availability of resources at a number of sites, the use of denuders is recommended for EC/OC measurements only and not a requirement for other chemical species. This means, however, that sampling artefacts do exist for a number of inorganic semi-volatile species, in particular when temperatures in the sampling system exceed 20 °C. This should be accounted for when reporting data to the World Data Centre for Aerosols.

High-volume and low-volume sampling lines are accepted in the context of GAW. For simplicity, it is suggested that a differencing technique be utilized to separate the coarse fraction from the fine fraction. Specifically, one filter should be run behind the 10 μm aerodynamic diameter cut inlet. A parallel filter should be run behind the inlet suitable for the fine fraction (i.e. 2.5 μm aerodynamic diameter at ambient relative humidity). While the second filter will yield the fine fraction, the difference between the two filters will yield the coarse fraction. For high-volume sampling, use of dichotomous samplers is an interesting alternative to differentiate fine and coarse aerosol fractions.

Low-volume samplers are less expensive than high-volume samplers. For the routine long-term aerosol measurements at GAW stations, it is recommended that up to three sets of 47 mm diameter filters be collected in parallel by low-volume samplers. If financial constraints are a limiting factor, the priorities for filter sampling are: (i) Teflon filters for gravimetric and ionic analyses; (ii) quartz-fibre filters for carbonaceous aerosol analyses with lines equipped with carbon denuders; and (iii) Teflon filters for elemental analyses. Each set would consist, ideally, of two filters, one for total mass below 10 μm diameter and one for the fine fraction. The separation would be achieved by running the filters behind the size-selective inlets. High-volume samplers are usually more expensive, and running more than one set of samplers in parallel is often impractical. For that reason, it is recommended that high-volume samplers be operated with quartz fibre filters for both inorganic and EC/OC analyses. It should be noted that denuders for high-volume samplers have only recently become commercially available.

There is no recommendation for determining sampling time as this will be highly site-dependent. In general, short sampling time (24 to 48 h) provides information that is more easily used in models and should be preferred over week-long sampling, even if discontinuous. Filters should be removed from the sampling unit shortly after collection and stored between 0 °C and 5 °C if analyses cannot be performed immediately. Regular blank samples should be taken in order to control for contamination. Ideally, such blanks are prepared by mounting filters into the sampling unit with the pump switched off. It is recommended that one blank be performed for every 10 samples.

For each GAW aerosol station, a list of core aerosol chemical measurements is strongly recommended: (i) mass; (ii) major ionic species; (iii) carbonaceous components, and (iv) dust aerosols.

The *mass concentration of atmospheric aerosols* is clearly a fundamental parameter in the GAW measurement programme. It is recommended that this be measured gravimetrically on Teflon filters. It is expressed in units of $\mu\text{g m}^{-3}$, where the volume is related to STP. Updated measurement guidelines on mass concentration measurements using gravimetric analysis of Teflon filters are provided in WMO (2011).

The tapered element oscillating microbalance (TEOM) has been widely used for aerosol mass measurements. The instrument provides continuous measurements and can produce high time resolution data. It is recommended that TEOM is used with Filter Dynamics Measurement System (FDMS) corrections to account for the loss of semi-volatile components of aerosols, and humidity.

A different type of instrument for continuous mass measurements, the β -meter, has also been commercialized. It operates on the principle of β -ray attenuation by a layer of aerosol. The β -ray source is usually ^{14}C or ^{85}Kr decay, and the attenuation can be calibrated with a known mass. Sampling can use individual filters or filter tapes, and the β -ray that passes through the filter is continuously monitored. The β -meters have the same inherent difficulties concerning volatilization as the TEOM. However, comparison with gravimetric methods usually produces reasonable agreement. Updated guidelines on the mass concentration measurements with beta attenuation (with the Met One Instruments model BAM-1020) are provided in WMO (2011).

The *concentration of major inorganic species* is one of the core pieces of information recommended for the GAW stations. Major ionic species include sulphate, nitrate, chloride, sodium, ammonium, potassium, magnesium and calcium. This selection is based on the fact that analytical procedures for these species have become well established. More importantly, under most atmospheric conditions, this set of ions is expected to account for a major part of the aerosol mass, and the measurements here are an important step towards mass closure of the aerosols. As mentioned

above, quantitative measurements of nitrate with a filter technique remain problematic and are associated with high uncertainties.

In the GAW Programme, it is recommended that analyses be performed using ion chromatography for the most cost-effective approach. The IC technique has the advantage of chemical speciation and relatively low cost per analysis, and has matured to the degree that the sensitivities for each ionic species, the cost and the maintenance are all reasonably well known. If IC is set up properly, all the recommended ionic species can be analysed in one single sample injection. Alternative analytical techniques exist but their use may introduce systematic differences among GAW stations. It is part of each laboratory's responsibility to document the equivalence of these alternative techniques, such as AAS or ICP-MS, with IC whenever they are used. Calibration of IC instruments is an integral part of every laboratory's standard operating procedures, and each laboratory must implement quality control procedures that guarantee the accuracy of calibrations. Recommendations for GAW are similar to those reported in WMO (2004a) for precipitation chemistry. In addition, protocols for filter extraction should be well documented.

Instruments are emerging on the scene for continuous and semi-continuous measurement of sulphate, nitrate, and organic carbon in aerosols. In particular, progress in aerosol mass spectrometry (AMS) has led to the development of instruments that can be used for monitoring purposes, providing quantitative measurements of the total mass and size distribution of non-refractory chemical composition in the submicron-size range. Simpler versions of AMS now exist commercially that can provide chemically speciated mass loadings and aerosol mass spectra with one-hour time resolution. No standard operating procedures endorsed by GAW / WMO exist yet although ACTRIS is operating a Central Facility targeting AMS that provides recommendations for long-term monitoring operations

http://www.actris.eu/Portals/46/Documentation/actris2/Deliverables/public/WP3_D3.3_M16.pdf?ver=2016-08-22-142809-857.

The *concentration of carbonaceous species* (with both elemental and organic fractions) is also one of the core pieces of information recommended at GAW stations. Carbonaceous species are still the least understood and most difficult to characterize of all aerosol chemical components. Total aerosol carbon mass can be divided into three fractions: inorganic carbonates, organic carbon (OC), and a third fraction ambiguously called elemental carbon (EC), black carbon (BC), soot or refractory carbon in the scientific literature with no clear definition of the terms. Recommendations for proper use of terminology for BC-related species have been proposed by Petzold et al. (2013) to clarify the terms used in atmospheric research; they recommend that the term "black carbon" be used only in a qualitative sense, and that terms related to the measurement technique be used when reporting quantitative results. According to this terminology, thermo-optical methods can be used to derive total carbon (TC) and OC/EC fraction in atmospheric aerosol filters. When using optical methods, the light-absorbing component is called equivalent black carbon (EBC), even though the optical method is not specific for carbon.

It is recommended that TC, OC and EC be measured in the GAW Programme, leaving out the relatively minor and difficult inorganic carbon component and the more complicated issue of organic carbon speciation. Sampling of aerosol carbonaceous materials is recommended using quartz filters, pre-fired at 350 to 400 °C for 2 h, and deployed at the same sampling frequency as the Teflon filters. The quartz filter can be analysed for TC using the thermal evolution technique. The mass concentration of the total carbon is obtained by thermal oxidation of the carbon, usually at 750 °C in the presence of a catalyst, to measurable carbon dioxide. Detection of the evolved carbon dioxide is done in one of two ways: either by reduction to methane in the presence of a catalyst, and then FID, or by direct detection by NDIR detectors.

The measurement of the TC components (OC and EC) is more difficult than the measurement of TC (Schmid et al., 2001). The distinction between the fractions is made by temperature-controlled volatilization/pyrolysis. This is followed by catalysed oxidation to CO₂ and detection by NDIR, or in some instruments, further catalysed reduction of the CO₂ to CH₄ and final detection by FID. There are different temperature-control programmes in use. At present, GAW recommends the use of one of three thermo-optical techniques: the Interagency Monitoring of Protected Visual Environments (IMPROVE) protocol (Chow et al., 1993; Chow et al., 2005), National Institute for Occupational Safety and Health (NIOSH) protocol (Birch and Cary, 1996) and European Supersites for

Atmospheric Aerosol Research (EUSAAR-2) protocol (Cavalli et al., 2010). Relatively good agreement is obtained between the IMPROVE, NIOSH and EUSAAR-2 protocols for TC determination, while they strongly differ for EC determination (Chow et al., 2001). Because EC represents a relatively small fraction of TC, OC determination by the three protocols is also comparable. It is accepted that the IMPROVE and EUSAAR-2 protocols are best suited for non-urban background sites, while the NIOSH is applied to samples from urban sites by the US Environmental Protection Agency (EPA). The use of IMPROVE (Watson et al., 2009) or EUSAAR-2 may therefore be preferred over NIOSH for global remote GAW stations. Use of the EUSAAR-2 and IMPROVE protocols for EC determination may lead to different results but should be preferred over any alternative techniques. Whenever an alternative approach is used for OC measurements, it is recommended that a periodic determination of OC using one of the thermo-optical methods be conducted so that results can be compared.

The use of optical methods for estimating EBC involves measuring the change in optical transmission of a deposit of particles on a filter (absorption) and applying a site-specific and instrument type-specific mass absorption efficiency to derive EBC. Two key assumptions are required to derive the equivalent black carbon mass concentration from light absorption measurements: (i) black carbon is the only species responsible for the aerosol light absorption, and (ii) the sampled black carbon has the same mass absorption efficiency as the standards used in laboratory calibrations of the absorption instrument. These assumptions can be evaluated by experimentally determining the mass absorption efficiency by simultaneously making light absorption measurements and EC measurements as described above. At sites where EC concentrations are not routinely determined on quartz-fibre filters, less frequent filter collections can be used to derive site- and season-specific values of the mass absorption efficiency. Thus, for GAW measurements of EBC, experimentally derived values of the mass absorption efficiency at a site are essential when estimating black carbon mass concentration from light absorption measurements.

The use of incandescent methods, such as single-particle soot photometers, or volatility techniques, such as volatility scanning mobility particle sizer (SMPS), can provide information on refractory material present in aerosol, but their use in monitoring activities at GAW stations remains problematic given the lack of standardized protocols and of consistent intercomparison with thermo-optical techniques.

Dust aerosols can be sampled relatively easily without the problems posed by more semi-volatile aerosol components such as organics and ammonium nitrate. For GAW stations, it is recommended that a multi-elemental analysis approach be used to determine the mineral dust component. Teflon filters should be analysed for at least four of the major crustal elements, aluminium (Al), silicon (Si), iron (Fe), titanium (Ti) and scandium (Sc), and the related elements, sodium (Na), magnesium (Mg), potassium (K) and calcium (Ca). No specific analytical technique is recommended as there is a good selection available, including proton-induced X-ray emission (PIXE), instrumental neutron activation analysis (INAA), X-ray fluorescence (XRF), AAS and ICP-MS. These techniques usually have high sensitivities for the crustal elements. Not all techniques can provide all the required elements, and depending on availability, a combination of two or more techniques may be necessary.

16.6.2 In situ measurements of aerosol radiative properties

The following aerosol radiative properties are needed for climate studies, all at multiple wavelengths across the visible spectrum:

- (a) Aerosol light extinction coefficient (σ_{ep}) and its two components (scattering and absorption).
- (b) Aerosol optical depth (δ , see 16.6.5), defined as the integral over the vertical column of the aerosol light extinction coefficient.
- (c) Aerosol single-scattering albedo (ω_0), defined as $\sigma_{sp}/(\sigma_{ap} + \sigma_{sp})$, describes the relative contributions of scattering and absorption to the total light extinction. Purely scattering aerosols (such as sulphuric acid) have values of 1, while very strong absorbers (such as black carbon) have values of about 0.3.

- (d) Radiative transfer models commonly require one of two integral properties of the angular distribution of scattered light (phase function): the asymmetry factor (g) or the upscatter fraction (β). The asymmetry factor is the cosine-weighted average of the phase function, ranging from a value of -1 for entirely backscattered light to $+1$ for entirely forward-scattered light. The upscatter fraction gives the fraction of sunlight scattered in the upwards direction (back to space), which depends on the solar zenith angle as well as the size distribution and chemical composition of the particles. It can be estimated from the hemispheric backscatter fraction (b).
- (e) Mass scattering efficiency for species i , (α_{si}), used in chemical transport models to evaluate the radiative effects of each chemical species forecast by the model. It is often calculated as the slope of the linear regression line relating the aerosol light scattering coefficient (σ_{sp}) and the mass concentration of the chemical species (though multiple linear regression is preferred, to deal with covariance of some chemical species). This parameter has units of $\text{m}^2 \text{g}^{-1}$.
- (f) Mass absorption efficiency for species i , (α_{ai}), used in chemical transport models to evaluate the radiative effects of each chemical species forecast by the model. It is often calculated as the slope of the linear regression line relating the aerosol light absorption coefficient (σ_{ap}) and the mass concentration of the chemical species (though multiple linear regression is preferred, to deal with covariance of some chemical species). This parameter has units of $\text{m}^2 \text{g}^{-1}$.
- (g) The functional dependence of components of the aerosol light extinction coefficient (σ_{ep} , σ_{sp} , σ_{ap}) on relative humidity, $f(\text{RH})$, expressed as a multiple of the value at a low reference relative humidity (typically $< 40\%$).

The aerosol light scattering coefficient is measured with an integrating nephelometer. Integrating nephelometers have been operated at baseline monitoring stations since the deployment of a four-wavelength instrument at the NOAA Mauna Loa Observatory in 1974. At present, there are about four dozen sites monitoring σ_{sp} routinely around the globe as part of the GAW global network. A few of these are operating single-wavelength units, but most are measuring σ_{sp} at three wavelengths. The multiwavelength integrating nephelometer TSI model 3563 operates at wavelengths of 450, 550 and 700 nm, and has the added feature of being able to measure σ_{sp} over two angular ranges: total scattering ($7\text{--}170^\circ$) and hemispheric backscattering ($90\text{--}170^\circ$, denoted as σ_{bsp}). The Aurora 3000 integrating nephelometer, manufactured by Ecotech, makes comparable measurements. While instruments do not exist for direct determination of g or β , the ratio $b = \sigma_{bsp}/\sigma_{sp}$ can be used to estimate either of these parameters (updated measurement guidelines are available in WMO (2011)). Simpler, less expensive and less sensitive one-wavelength instruments are also commercially available. These instruments can provide useful information on the aerosol light scattering coefficient at regional sites where aerosol loadings allow the use of a less sensitive instrument.

Instruments capable of high time-resolution determination of the aerosol light absorption coefficient are commercially available. They are based on the rate of change of transmission through a fibre filter as particles are deposited on the filter. Calibration of these filter-based methods is difficult but required because the relationship between the change in light transmission and aerosol absorption optical depth on the filter depends on many factors, including the particular filter medium and the light-scattering nature of the particles.

One instrument in common use is the aethalometer. Originally, this instrument was calibrated in terms of an equivalent mass of black carbon rather than the fundamental property that provides the instrumental response: aerosol light absorption. Early models of the aethalometer have a very broad wavelength response, while newer versions offer narrowband measurements at multiple wavelengths.

Another commercial, filter-based instrument for determining σ_{ap} is the particle soot absorption photometer (PSAP) that measures laboratory aerosols with different single-scattering albedos, using a calibration standard based on the difference between σ_{ep} , measured with a long-path extinction cell, and σ_{sp} , measured with an integrating nephelometer. Updated measurement guidelines for PSAP instruments are provided in WMO (2011).

Yet another filter-based instrument is the multi-angle absorption photometer (MAAP). The MAAP uses a different optical configuration than the aethalometer and PSAP, with measurements of the filter reflectivity at two different angles in addition to the filter transmission measurement. The two reflectivity measurements allow correction for multiple scattering processes involving the deposited particles and the filter matrix. This approach eliminates the need for a correction scheme based on independent measurements of the aerosol light scattering coefficient. The MAAP operates at a wavelength of 670 nm. Updated measurement guidelines for MAAP instruments are provided in WMO (2011).

Recent improvements in a different approach to determining the aerosol light absorption coefficient, called photoacoustic spectroscopy, offer a promising alternative to filter-based methods. Although not as sensitive, the photoacoustic method allows determination of the aerosol light absorption coefficient while the particles are suspended in air, eliminating the artefacts introduced by depositing the particles on a filter. The photoacoustic method can be used in regions where light absorption levels are moderately high, and as a calibration standard for filter-based instruments.

16.6.3 Particle number concentration and size distribution

The condensation particle counter (CPC) is used to measure particle number concentration for particles greater than 10 nm. The technology is well established and is commercially available from different manufacturers. Condensation nuclei (CN) can be detected after the condensation of water or other condensable vapour (often an alcohol such as butanol) from a supersaturated atmosphere onto the particle. The super-saturation in condensation nuclei counters (CNCs), which are also known as condensation particle counters (CPCs), is typically quite large, about 150 %, allowing the detection of particles as small as a few nanometres in diameter. The CPC is robust and designed for long-term operations. However, it needs regular checks on-site to quality-assure the measurement, especially in terms of cleaning the saturator and the optics. Additionally, annual calibrations of the CPC against a reference instrument are required. If the counting efficiency is not controlled, CPC performance might drift with time causing an unnoticed bias of up to several tens of percent. Measurements of particle number size distribution complement CN counters. Many commercial instruments are now available for both the fine and the coarse modes. They utilize a wide range of physical principles to classify particles according to size. Some of the better-known approaches use the electrical mobility of particles, aerodynamic size, or optical size determined by light scattering. Mobility particle size spectrometers (MPSS) measure the particle number size distribution of the submicrometer size range from approximately 10 to 800 nm. The technology is well established and is commercially available, but there are also custom designed measurement systems. A MPSS is robust and designed for long-term operations although it requires regular checks on-site to quality-assure the measurement. Additionally, annual or biannual calibrations of both MPSS and CPC against reference instruments are required. If the MPSS instrument performance is not controlled, MPSS might drift with time causing unnoticed biases in both particle sizing and particle number concentration of up to several tens of percent.

For the upper accumulation and coarse mode size ranges, optical and aerodynamic particle size spectrometers (OPSS & APSS) are employed. The technology of the APSS is well established and is commercially available. Comparison results of these instruments can be found at <http://www.wmo-gaw-wcc-aerosol-physics.org/files/internal-actris-report-intercomparison-workshop-apss-2014.pdf>. The APSS is robust and designed for long-term operations. There are several different OPSS on the market, some of which could be used for long-term measurements. OPSS include a number of relatively small, low-cost instruments utilizing laser diodes that have to be operated carefully to deliver reliable quantitative measurements. Both APSS and OPSS need regular sizing checks on-site to quality-assure the measurement. Additionally, annual or biannual calibrations against traceable standards and reference instruments are required. If OPSS or APSS performance is not controlled, OPSS or APSS instruments might drift with time causing an unnoticed bias of up to several tens of percent.

16.6.4 Cloud condensation nuclei

Cloud condensation nuclei (CCN) measurements are made to determine the concentration and establish climatology of those particles that have the potential to produce cloud droplets at supersaturations typical of natural clouds, that is, less than about 1 %. Past CCN measurements in the

GAW Programme have been made predominantly using static thermal-gradient chambers, which are well suited to relatively low-frequency sampling and low-resolution (differential) CCN spectrum determination. Instruments utilizing continuous flow now have established technology and are commercially available, becoming the recommended technical approach being deployed at a growing number of GAW stations. The CCN counter (CCNC) designed for long-term operations. However, it needs regular calibrations with a standard aerosol (every six month) on-site to quality-assure the measurement. Additionally, annual or biannual calibrations of the CCNC against a reference instrument are required. If CCNC performance is not controlled, CCNC flow might change with time causing an ill-defined supersaturation in the instrument and a bias in the measured number concentration. Because of the complexity of the measurement, it is recommended that CCN determinations be undertaken at GAW stations with more highly developed aerosol programmes.

16.6.5 Aerosol optical depth

Aerosol optical depth (AOD) is retrieved from observations of atmospheric spectral transmission. The solar spectral irradiance I at a given wavelength can be expressed as:

$$I = I_0 \exp(-m\delta) \quad (16.3)$$

with I_0 being the extraterrestrial (top of the atmosphere) irradiance of the sun, m the optical air-mass and δ the total optical depth. The optical air-mass equals 1 for a vertical path through the atmosphere and is roughly proportional to $1/\cos z$, with z being the zenith angle of the sun during the observation. The total optical depth δ at a given wavelength is composed of several components, such as scattering by gas molecules δ_R (Rayleigh scattering), extinction by aerosol particles δ_A , absorption of trace gases δ_G (ozone, nitrogen dioxide, etc.), and possible cloud contamination. Thus, the AOD can be obtained from the total optical depth by subtracting modelled estimates of the other components $\delta_A = \delta - \delta_R - \delta_G$.

Because AOD is essentially a difference between two larger numbers, it is sensitive to small calibration errors and, to a minor degree, to the methods chosen to model the other components. A traceable calibration uncertainty of 1.5%, corresponding to an uncertainty of 0.015 optical depths at unit optical air-mass, should be maintained for AOD observations (WMO, 2005).

Wavelengths and bandpasses specifically for AOD that are largely free of variable extinction components (water vapour and NO_x) and strong ozone extinction have been recommended in WMO (2003). The Baseline Surface Radiation Network (BSRN) and GAW-PFR (a network of aerosol optical depth observations with precision filter radiometers) are using four AOD channels at 368, 412, 500 and 862 nm. While some other networks have selected different wavelengths based on their specific needs (validation of satellite sensors, modelling efforts), measurements at 500 ± 3 nm and 865 ± 5 nm are typically available in most networks.

Measurements of the solar spectral irradiance are traditionally taken by sun-pointing radiometers (sun photometers) mounted on a two-axis solar tracker, with a sampling rate of once every minute to allow for objective quality control and cloud filtering algorithms. Homogeneous quality control is more difficult to achieve with handheld sun photometers.

Rotating shadow-band filter radiometers measure global and diffuse spectral irradiance at several wavelength bands. Direct normal irradiance obtained as the difference between global and diffuse radiation, normalized by the solar zenith angle, can be used to retrieve AOD in the same way as with sun photometers.

More advanced instruments like sky-scanning radiometers can be used to infer additional column aerosol optical properties, including size distribution, single-scattering albedo or phase function, through sophisticated mathematical inversion models.

Sun photometers and shadow-band and sky-scanning radiometers are commercially available from several manufacturers. Centralized data evaluation and calibration services⁶ are offered for standardized instruments by global networks, such as the Aerosol Robotic Network (AERONET), GAW-PFR or SKYNET (WMO, 2005).

16.6.6 GAW aerosol lidar

The basic lidar principle is the following: a laser pulse is transmitted into the atmosphere where it encounters gas molecules and particles; a small amount of this energy is backscattered in the direction of the receiver system, typically a telescope, and transferred to a photodetector such as a photomultiplier. The resulting electrical signal is proportional to the optical power received, which depends on the presence, range and concentration of atmospheric scatterers and absorbers in the light path volume. Lidar techniques are able to characterize atmospheric aerosols in terms of vertical profiles of extinction and backscatter coefficients, lidar ratio, optical depth and microphysical properties such as shape, refractive index and size distribution, on timescales as short as minutes and vertical scales as short as metres.

Lidar observations are much more powerful if used in coordinated networks. Lidar networks are fundamental to studying aerosols on a large spatial scale and to investigating transport and modification phenomena. There are several research lidar networks that contribute to GAW: the Asian Dust and Aerosol Lidar Observation Network (AD-NET), Latin American Lidar Network (LALINET or ALINE), Commonwealth of Independent States Lidar Network (CIS-LINET), European Aerosol Research Lidar Network (EARLINET), Micro Pulse Lidar Network (MPLNET), NDACC, and the NOAA Cooperative Remote Sensing Science and Technology (CREST) Lidar Network. These networks are coordinated within the GAW Aerosol Lidar Observation Network (GALION; WMO, 2008c).

Several different lidar techniques exist, depending on the specific instrument design and mainly on the specific laser-atmosphere scattering process.

Elastic backscatter lidar

This is the simplest type of aerosol lidar: the backscattered wavelength is identical to the transmitted wavelength, and the magnitude of the received signal at a given range depends on the backscatter cross-section of scatterers along the path to that range. Typical operating wavelengths are 355, 532 and 1 064 nm. The typical product of a backscatter lidar is the vertical profile of the aerosol backscatter coefficient obtained assuming a lidar ratio, i.e. the extinction-to-backscatter ratio, that is mostly constant throughout the profile and usually derived from an existing climatology obtained with measurements from a Raman lidar, described later. In this sense, it is necessary to underline that without an a priori assumption about the lidar ratio, this kind of lidar system cannot provide quantitative aerosol backscatter data.

Depolarization lidar

These are elastic backscatter lidars equipped with channels for the detection of the two parallel and cross-polarized components of the backscattered radiation. This provides quantitative information about particle shape, strongly contributing to aerosol typing as well as to the identification of thin clouds contaminating the profiles. A depolarization channel allows discrimination of volcanic ash and other aerosol particles. Typical operating wavelengths are 355 and 532 nm. Depolarization lidar systems need accurate calibration.

⁶ For more details, see the World Optical Depth Research and Calibration Centre website (at <http://www.pmodwrc.ch/worcc/index.html>).

Raman lidar

The Raman lidar technique operates by measuring the inelastic Raman scattering by a specific gas. The Raman backscattered radiation from molecular nitrogen (or oxygen) is typically used for retrieving the vertical profile of the aerosol extinction coefficient that, coupled with the elastic scattering collected at the same emission wavelength, also provides the vertical profile of the aerosol backscatter coefficient without assuming a lidar ratio. Typical operating wavelengths are 355 and 532 nm. Most of the existing Raman lidar instruments are also equipped with a depolarization channel providing data on the particle linear depolarization ratio. Advanced multiwavelength Raman aerosol lidar techniques have been demonstrated to be the only technique capable of providing range-resolved aerosol microphysical properties. Moreover, rotational Raman lidar systems can be designed for optimizing extinction measurements in daytime conditions.

High spectral resolution lidar

The high spectral resolution lidar (HSRL) technique provides calibrated measurements of aerosol optical depth, extinction and backscatter. Measurements are computed from ratios of the particulate scattering to the measured molecular scattering. This provides absolute calibration and makes the calibration insensitive to dirt or precipitation on the output window. A very narrow, angular field-of-view reduces contamination from spurious sources, like multiple scattering contributions. The small field-of-view, coupled with a narrow optical bandwidth, nearly eliminates noise due to scattered sunlight, improving the signal-to-noise ratio during daytime operations.

Ceilometers

Ceilometers are basically elastic backscatter lidars that employ a diode laser source emitting at infrared wavelengths (typically 905 or 1 064 nm) using a low energy but a high repetition rate (in units of μJ of energy per pulse and kHz for the rate) and that detect the elastic backscattered radiation by clouds and precipitation. Ceilometers are a self-contained, turnkey, surface-based, active, remote-sensing device designed to measure cloud-base height and potentially the backscatter signals by aerosols. Ceilometers can provide qualitative information about aerosol vertical distribution. Generally, older and typically less powerful instruments are barely able to detect aerosol layers in the atmosphere, while newer instruments can be useful for volcanic ash/dust detection and ash/dust plume tracking.

All of these lidar techniques can provide data products suitable for monitoring the spatial and temporal distribution of aerosol up to the upper troposphere/lower stratosphere region and can characterize them from a dynamical and microphysical point of view. The main lidar limitation is related to the presence of rain, dense fog and thick clouds (optical depths larger than 2–3) that do not allow monitoring of the atmosphere above the cloud-base region. The altitude range covered by lidars is limited at the bottom by the overlap height (altitude where there is a full overlap between the transmitter and the receiver), typically about 250–500 m above the ground level but can be up to 2 km above the ground depending on the specific design. The maximum altitude range strongly depends on the laser power and optical design, reaching 25–30 km for high-power systems. It is difficult to provide general estimates of the accuracy of the different lidar products because these are system specific and also depend on the prevailing meteorological conditions. On average, uncertainties for extinction and backscatter coefficient are about 20 % (in the case of Raman lidar or HSRL). The retrieval of microphysical properties is possible only if optical data have uncertainties lower than 20 %–30 %.

Aerosol lidar products (see Table 16.4 for more details):

- (a) Geometrical properties:
 - (i) Layer identification (top, bottom and centre of mass);
- (b) Profiles of optical properties:
 - (i) Extensive optical parameters: aerosol backscatter coefficient (β_a), aerosol extinction coefficient (α_a);

- (ii) Intensive optical parameters: lidar ratio, particle linear depolarization ratio (δ_a), Ångström backscatter-related exponent (Å_β), Ångström extinction-related exponent (Å_α);
- (c) Optical properties in the identified layer:
 - (i) Integrated backscatter, aerosol optical depth;
 - (ii) Mean intensive optical parameters (lidar ratio, particle linear depolarization ratio, Ångström backscatter-related exponent, Ångström extinction-related exponent);
- (d) Aerosol typing classification;
- (e) Mass concentration estimate;
- (f) Microphysical properties retrieved.

For Members' Review

**Table 16.4. Lidar products related to specific surface-based lidar techniques
(note that (d) = daytime only)**

TABLE: Table horizontal lines

<i>Surface-based lidar techniques</i>	<i>Geometrical properties</i>	β_a	α_a	<i>Lidar ratio</i> ^a	<i>AOD</i>	\hat{A}_β	\hat{A}_α	<i>Type</i> ^b	<i>Microphysical properties</i>
Ceilmeter ^c	✓	✓ ^d							
Ceilmeter + sun photometer	✓	✓	✓(d) ^e		✓(d)				
Ceilmeter + sun photo. + depolarization lidar	✓	✓	✓(d) ^e		✓(d)			✓ (limited)	
1-wavelength (1- λ) backscatter lidar	✓	✓							
1- λ backscatter lidar + sun photometer	✓	✓	✓(d) ^e		✓(d)				
1- λ backscatter lidar + sun photo. + depolarization lidar	✓	✓	✓(d) ^e		✓(d)			✓(d) (limited)	
Multiwavelength (m- λ) ^f backscatter lidar	✓	✓					✓		
M- λ ^f backscatter lidar + sun photometer	✓	✓	✓(d) ^e		✓(d)	✓(d) ^e	✓		✓(d) ^e
M- λ ^f backscatter lidar + sun photo. + depolarization lidar	✓	✓	✓(d) ^e		✓(d)	✓(d) ^e	✓	✓	✓(d) ^e
1- λ Raman lidar/HSRL	✓	✓	✓ ^g	✓ ^g	✓ ^g			✓ ^g (limited)	
1- λ Raman lidar/HSRL + sun photometer	✓	✓	✓ ^g	✓ ^g	✓ ^g	✓(d) ^e	✓(d) ^e	✓ (limited)	✓(d) ^e
1- λ Raman lidar/HSRL + sun photo. + depolarization lidar	✓	✓	✓ ^g	✓ ^g	✓ ^g	✓(d) ^e	✓(d) ^e	✓	✓(d) ^e
M- λ ^f Raman lidar	✓	✓	✓ ^g	✓ ^g	✓ ^g	✓ ^g	✓	✓ ^g	✓ ^g
M- λ ^f Raman lidar + sun photometer	✓	✓	✓ ^g	✓ ^g	✓ ^g	✓ ^g	✓	✓	✓ ^g
M- λ ^f Raman lidar + sun photo. + depolarization lidar	✓	✓	✓ ^g	✓ ^g	✓ ^g	✓ ^g	✓	✓	✓ ^g

Notes:

- a From two independent measurements
- b Identification of scattering type (aerosol particles, cloud droplets, ice crystals, some aerosol type information)
- c A ceilometer is a single-wavelength, low-power lidar, with lower signal-to-noise ratio.
- d If calibrated
- e Estimate only
- f $m > 2$
- g Most Raman lidar systems operate during night-time. Some 24-h Raman lidar systems exist and their operability has been proved. However, few systems nowadays operate Raman channels during daytime; HSRL is independent of day-time.

16.7 NATURAL RADIOACTIVITY

The global distributions of the source/sink terms of the naturally occurring radionuclides (^7Be , ^{10}Be , ^{210}Pb and ^{222}Rn) and the anthropogenic radionuclides (^{85}Kr) are reasonably well known. ^7Be and ^{10}Be are produced by cosmic-ray interactions in the upper troposphere and lower stratosphere. ^{222}Rn is exhaled from the Earth's land surface as a result of uranium decay in soil. ^{210}Pb is produced in the lower troposphere from the decay of ^{222}Rn . Most of the ^{85}Kr in the atmosphere is released during nuclear fuel reprocessing. Atoms of ^7Be , ^{10}Be and ^{210}Pb attach themselves to sub-micron-size aerosol particles, and therefore act as aerosol-borne tracers in the atmosphere. ^{222}Rn and ^{85}Kr , which are chemically and physically inert, act as noble gases in the atmosphere.

Measurements of radionuclides are not a priority area within the GAW Programme. Some general recommendations can be found in WMO (2001) and WMO (2004b).

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ANNEX. GAW CENTRAL FACILITIES

List of GAW central facilities and host institutions as of September 2013; the World Central Facilities have assumed global responsibilities, unless indicated.

TABLE: Table horizontal lines

Variable	Quality Assurance/Science Activity Centre	Central Calibration Laboratory (host of primary standard)	World Calibration Centre	Regional Calibration Centre	World Data Centre
CO_2	JMA (A/O) ^a	NOAA-ESRL	NOAA-ESRL (round robin) Empa (audits)		JMA
Carbon isotopes		MPI-BGC			JMA
CH_4	Empa (Am, E/A) ^a JMA (A/O) ^a	NOAA-ESRL	Empa (Am, E/A) ^a JMA (A/O) ^a		JMA
N_2O	UBA	NOAA-ESRL	IMK-IFU		JMA

CFCs, HCFCs, HFCs					JMA
SF ₆		NOAA-ESRL	KMA		JMA
H ₂		MPI-BGC			JMA
Total ozone	JMA (A/O) ^a	NOAA-ESRL, ^b EC ^c	NOAA-ESRL, ^b EC ^c	BoM, ^b NOAA-ESRL, ^b EC, ^e DLR ^f IZO, ^c JMA, ^b MOHp, ^b MGO, ^d OCBA, ^b SAWS, ^b SOO-HK ^b	
Ozonesondes	IEK-8	IEK-8	IEK-8		EC
Surface ozone	Empa	NIST	Empa	OCBA	JMA, NILU ^g
Precipitation chemistry	NOAA-ARL	ISWS	ISWS		NOAA-ARL
CO	Empa	NOAA-ESRL	Empa		JMA
VOCs	UBA	NPL, NIST, KRISS	IMK-IFU		JMA
SO ₂					JMA
NO _x	UBA		IEK-8 (NO _x)		JMA
Aerosol	UBA (physical properties)		TROPOS (physical properties)		NILU, ^f DLR ^g
Optical depth		PMOD/WRC ^h	PMOD/WRC		NILU
UV radiation			PMOD/WRC	NOAA-ESRL (Am), ^a EUVC/PMOD (E) ^a	EC
Solar radiation		PMOD/WRC	PMOD/WRC		MGO

Notes:

a Am = Americas; E/A = Europe and Africa; A/O = Asia and the South-West Pacific

b Dobson

c Brewer

d Filter instruments

e Surface-based

f Space-based

g Since January 2016

h Precision filter radiometers

Host institutions

TABLE: Table as text NO space

BoM	Bureau of Meteorology, Melbourne, Australia
BSRN	Baseline Surface Radiation Network, Federal Institute of Technology (ETH), Zürich, Switzerland
DLR	German Aerospace Centre, Oberpfaffenhofen, Wessling, Germany

EC	Environment Canada, Toronto, Canada
Empa	Swiss Federal Laboratories for Materials Testing and Research, Dübendorf, Switzerland
EUVC	European Ultraviolet Calibration Centre (see PMOD/WRC)
IEK-8	Institute of Energy and Climate Research: Troposphere (IEK-8), Research Centre Jülich GmbH, Jülich, Germany
TROPOS	Institute for Tropospheric Research, Leipzig, Germany
IMK-IFU	Karlsruhe Institute of Technology (KIT), Institute of Meteorology and Climate Research – Atmospheric Environmental Research, Garmisch-Partenkirchen, Germany
ISWS	Illinois State Water Survey, Champaign, IL, United States
IZO	Izaña Observatory, Tenerife, Spain
JMA	Japan Meteorological Agency, Tokyo, Japan
KMA	Korea Meteorological Administration, Seoul, the Republic of Korea
KRISS	Korea Research Institute of Standards and Science, Daejeon, the Republic of Korea
MGO	A.I. Voeikov Main Geophysical Observatory, Russian Federal Service for Hydrometeorology and Environmental Monitoring, St. Petersburg, the Russian Federation
MPI-BGC	Max Planck Institute for Biogeochemistry, Jena, Germany
MOHp	Meteorologisches Observatorium Hohenpeissenberg, Hohenpeissenberg, Germany
NOAA-ARL	National Oceanic and Atmospheric Administration, Air Resources Laboratory, Silver Spring, MD, United States
NOAA-ESRL	National Oceanic and Atmospheric Administration, Earth System Research Laboratory, Global Monitoring Division, Boulder, CO, United States
NILU	Norwegian Institute for Air Research, Kjeller, Norway
NIST	National Institute of Standards and Technology, Gaithersburg, MD, United States
NPL	National Physical Laboratory, Teddington, Middlesex, United Kingdom
OCBA	Observatorio Central de Buenos Aires, Argentina
PMOD/WRC	Physikalisch-Meteorologisches Observatorium Davos/World Radiation Centre, Davos, Switzerland
SAWS	South African Weather Service, Pretoria, South Africa
SOO-HK	Solar and Ozone Observatory, Hradec Kralove, Czech Republic
UBA	German Environmental Protection Agency, Berlin, Germany

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